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# Influence of landscape management practices on urban greenhouse gas budgets



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## Abstract

**Background:** With a lack of United States federal policy to address climate change, cities, the private sector, and universities have shouldered much of the work to reduce carbon dioxide (CO<sub>2</sub>) and other greenhouse gas emissions. This study aims to determine how landcover characteristics influence the amount of carbon (C) sequestered and respired via biological processes, evaluating the role of land management on the overall C budget of an urban university. Boston University published a comprehensive Climate Action Plan in 2017 with the goal of achieving C neutrality by 2040. In this study, we digitized and discretized each of Boston University's three urban campuses into landcover types, with C sequestration and respiration rates measured and scaled to provide a University-wide estimate of biogenic C fluxes within the broader context of total University emissions.

**Results:** Each of Boston University's three highly urban campuses were net sources of biogenic C to the atmosphere. While trees were estimated to sequester  $0.6\pm0.2$  kg C m<sup>-2</sup> canopy cover year<sup>-1</sup>, mulch and lawn areas in 2018 emitted C at rates of  $1.7\pm0.4$  kg C m<sup>-2</sup> year<sup>-1</sup> and  $1.4\pm0.4$  kg C m<sup>-2</sup> year<sup>-1</sup>, respectively. C uptake by tree canopy cover, which can spatially overlap lawn and mulched landcovers, was not large enough to offset biogenic emissions. The proportion of biogenic emissions to Scope 1 anthropogenic emissions on each campus varied from 0.5% to 2%, and depended primarily on the total anthropogenic emissions on each campus.

**Conclusions:** Our study quantifies the role of urban landcover in local C budgets, offering insights on how landscaping management strategies—such as decreasing mulch application rates and expanding tree canopy extent—can assist universities in minimizing biogenic C emissions and even potentially creating a small biogenic C sink. Although biogenic C fluxes represent a small fraction of overall anthropogenic emissions on urban university campuses, these biogenic fluxes are under active management by the university and should be included in climate action plans.

**Keywords:** Climate action plan, Urban carbon cycling, Landscaping, Biogenic fluxes, Soil respiration, Nature-based solution

## **Background**

With a lack of United States federal policy to address climate change, cities, the private sector, and universities have shouldered much of the work to reduce carbon dioxide (CO<sub>2</sub>) and other greenhouse gas emissions.

These cities—including their colleges, corporations, and faith groups—represent 155 million people [1]. While  $\mathrm{CO}_2$  emissions from fossil fuel combustion have been a focus for local climate action plans, biogenic fluxes and the broader carbon (C) cycle in urban ecosystems have received less attention [2]. Many urban greenhouse gas models ignore urban biogenic fluxes due to the misconception that urban areas have negligible amounts of vegetation and soil [3, 4]. Recently, an increasing number of studies have evaluated biogenic fluxes and their

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contribution to C storage in ecosystems by considering spatially heterogeneous vegetation [5] and variability in soil management practices [6].

Explicitly accounting for biogenic fluxes in urban ecosystems is critical for quantifying urban C fluxes [7], as ignoring biological contributions can impact the accuracy of CO<sub>2</sub> measurements via atmospheric monitoring [8] and overall reporting of total CO<sub>2</sub> fluxes in urban ecosystems [4]. Climate action plans tend to focus on emissions levers like transportation policies aimed at reducing automobile dependence and increasing mass transit ridership [9]. Purchasing renewable energy and setting regulations to improve building energy efficiency as a means of reducing greenhouse gas emissions are also common mechanisms of climate action plans [10]. However, it is critical to understand how green infrastructure choices might create nature-based solutions to offset CO2 emissions [11] or whether they represent an unquantified additional source of emissions due to landscape management choices.

On a smaller scale, many institutions—particularly universities—have developed sustainability plans to leverage their resources and expertise to reduce their net greenhouse gas emissions, influencing the CO<sub>2</sub> reduction goals of their respective cities. Boston University has pledged to achieve C neutrality by 2040, ten years before the City of Boston's target year of 2050, and has a comprehensive plan to reduce CO<sub>2</sub> emissions through renewable energy and increased energy efficiency [12]. Similar to other institutional actors, Boston University included C sequestration in non-urban tracts of forest that are not on the campus, but owned by the University, in their climate action plan. However, design and management of green infrastructure on urban campuses themselves may help to mitigate some of the fossil fuel fluxes and to potentially reduce energy demand via canopy shading and wind breaks [13–15]. However, mitigation of fossil fuel emissions using green infrastructure was not explicitly included in Boston University's climate action plan, and consequently existing nature-based solutions were not considered in this study. While food waste, human respiration, and many other factors also influence the biogenic C balance [16], this paper focuses on biogenic C fluxes in plants and soils of urban ecosystems across three highly urban campuses owned by Boston University.

Ecosystem C exchange includes uptake of C through photosynthesis and losses through respiration, with the difference between these two processes resulting in either a biological sink (uptake) or source (release) of atmospheric  $CO_2$ . Soil respiration contributes about 70% of respiratory losses in  $CO_2$  in temperate forests [17], while  $CO_2$  fluxes from soil respiration in lawn and landscaped landcover in urban areas can be the same

order of magnitude as fossil fuel emissions, especially in intensively landscaped environments such as suburbs [6]. Management choices to amend the soil (e.g. addition of compost and mulch) can more than double CO<sub>2</sub> respiration rates relative to nearby rural forests [6]. Increased rates of soil respiration due to mulch application, for example, mean that landscaping "beautification" efforts can have significant effects on the rate and magnitude of C emitted from soils. Landscaping maintenance activities also have direct fossil fuel emissions associated with activities such as crew and product transportation, lawn mowing, leaf blowing, and fertilizer application [18], as well as the operation of bucket trucks, chainsaws, and wood chippers during tree maintenance [19]. Landcover management decisions in urban areas therefore have the potential to greatly influence the direction and magnitude of biogenic C cycling in urban ecosystems.

Similar to soil respiration, C sequestration in trees can be greater in urban ecosystems compared to adjacent rural areas. Briber et al. [20] found that urban tree growth rates accelerate up to twice those in rural forests following urbanization, suggesting that  $CO_2$  uptake in urban trees plays a significant role in the urban C cycle. However, the shorter lifespan of urban trees compared to rural trees can affect the ecosystem's ability to effectively store C on a large scale [21].

Few studies have quantified C fluxes through urban ecosystems using bottom-up estimates of landcover type and flux rate. Hardiman et al. [4] estimated biogenic C fluxes for Massachusetts using a 30 m resolution biomass map coupled with the Vegetation Photosynthesis and Respiration Model. Other bottom-up approaches assessed the effects of human presence on soil C content [22] and compared estimates of emissions from energy consumption to local biogenic uptake measurements [23]. Additional modeling approaches have used established remote sensing methods—such as Light Detection and Ranging (LiDAR) and multi- and hyperspectral images—to map urban forest structure and estimate C storage [3, 24, 25].

Our study uses in situ measurements and meter-scale landcover classification data to quantify C fluxes in terrestrial ecosystems on Boston University's urban campuses to assess the role of biogenic C fluxes within the broader C budget. We focused primarily on C loss from mulch and lawn respiration and net C uptake by trees and grass, as well as anthropogenic CO<sub>2</sub> emissions associated with heating and electricity (referred to as Scope 1 and Scope 2 emissions, respectively). This study fills an important gap in knowledge of C management research, as urban areas historically have not been major focal points for ecologists examining the role of biogenic fluxes in the C cycle [26].

## **Materials and methods**

#### Site description

The greater Boston area is the most populous metropolitan area in Massachusetts and the tenth largest metropolitan area in the United States [27]. Boston has a temperate climate, with mean winter and summer temperatures of - 0.1 °C and 21.7 °C, respectively; mean annual precipitation is approximately 1112 mm [28]. Boston University owns land on five campuses: Charles River, Medical, Wheelock, Sargent Center, and Tanglewood Campuses. Boston University's main campus, the Charles River Campus (74 ha), lies parallel to the Charles River in the City of Boston (Additional File 1). Boston University owns two additional urban campuses in Boston; the Medical Campus (32 ha) in the South End and the Wheelock Campus (3.2 ha) in the Fenway neighborhood of Boston (Additional File 1). Outside of Boston, the University owns the Sargent Center (302 ha), a conference center and Nature's Classroom in Hancock, New Hampshire, and the Tanglewood Campus (26 ha), home to the Boston University Tanglewood Institute in Lenox, Massachusetts, both of which are rural.

In early June 2017, ten field plots ranging from 179 to 3963 m<sup>2</sup> (Table 1) were established on the Charles River Campus. Plot locations were selected based on patch sizes and to be representative of the two dominant vegetated landcover types present on the three urban Boston University campuses: lawn and mulched landscapes. Each plot was dominated by ornamental, non-native hardwood trees common to northeastern United States cities. All trees were intentionally planted and are actively managed by the University landscaping crew.

## Soil respiration and tree growth

We installed two or three 20.2 cm-diameter polyvinyl chloride (PVC) collars per plot in June 2017 to measure

CO<sub>2</sub> efflux across the two landcover types (Table 1). Plots with more than one landcover type present had one soil collar installed in each landcover type. We had a total of nine collars on lawns and nine in mulched areas. Consistent with Decina et al. [6], soil collars were left to equilibrate for 2–3 weeks in order to avoid large pulses in CO<sub>2</sub> efflux from roots severed during installation. CO<sub>2</sub> flux was measured using a 20-cm chamber soil CO2 efflux system (LiCOR-8100A) from 05 July to 31 October 2017 and 8 May 2018 to 29 November 2018, with an additional measurement made on 28 January 2019. Because effectively all landscape management occurs during the growing season (April-October), observations made across the entire growing season ensured that measurements represented a "typical" year of landscape management on Boston University's urban campuses. During the winter months, deciduous tree growth is effectively zero and soil respiration rates are extremely low with very cold and potentially snow covered soils. Soil respiration measurements were made weekly between late morning and midafternoon in a randomized order across plots to account for variation across time of day.

To measure rates of tree growth, dendrometer bands were installed on hardwood trees across the ten plots (n=73 trees total) in June 2017. Dendrometer bands were placed at 1.37 m above the ground to measure changes in tree diameter over time. Tree growth was measured monthly from 10 May through 30 November 2018. Measurements were made using digital calipers. In 2018, dendrometer band and soil flux measurements continued on eight out of ten plots, as one plot (Plot 5) was removed from sampling due to peeling tree bark and another (Plot 8) was removed entirely due to paving in part of the plot. A final total of 50 trees were measured monthly for growth.

Table 1 Areas of plots on the Boston University campus. Number of trees with dendrometer bands and number of soil collars are listed for each plot in 2017, with the number of samples for 2018 in parentheses

Plot number	Area (m²)	Cover types included	Number of trees with dendrometer bands	Number of soil collars	
1	1427.09	Lawn, mulch	8 (3)	2 (1)	
2	870.11	Lawn	8 (7)	2 (2)	
3	179.25	Mulch	8 (8)	2 (2)	
4	389.37	Lawn, mulch	8 (8)	2 (2)	
5	3101.81	Lawn	8 (0)	2 (0)	
6	2175.38	Lawn, mulch	8 (8)	3 (2)	
7	3962.99	Lawn, mulch	8 (8)	2 (1)	
9	182.93	Mulch	4 (4)	0 (0)	
11	544.36	Lawn, mulch	5 (4)	3 (3)	

#### Landcover data processing

Shapefiles with parcel boundaries were used to define the study areas across the five campuses, with the exception of the Wheelock Campus (acquired in 2018), which was manually digitized using Google Earth (Version 7.3). All parcels were projected onto the Google Earth base map, and 13 landcover classes were manually digitized with a minimum mapping unit of roughly 1 m<sup>2</sup>: artificial impervious surface (i.e., track, AstroTurf, rubber rail crossings), bare soil, building, concrete, forested, gravel/pavers/ brick/packed dirt road, landscaped garden, lawn, mulch, parking lot, paved, water, and thinned forest (Fig. 1). An overlaying tree canopy class was created to identify and quantify where tree canopy covers the Boston University campuses (Figs. 1, 2). Google Earth satellite imagery was used primarily from April 2017 and was supplemented with imagery from June 2017 and October 2018. When possible, landcover classifications were verified using Google Earth's Street View feature or visited in-person. Canopy cover polygons were drawn using leaf-on imagery from June 2017, although imagery from October 2018 was also used to delineate Tanglewood Campus canopy cover. The digitization process resulted in 5,560 unique polygons across all five campuses.

Using ArcGIS Pro (Version 2.0.1), each polygon was converted into a shapefile and projected onto a map using the North American Datum 1983 StatePlane Massachusetts Mainland 2001 projected coordinate system. The geodatabase produced by ArcGIS was read into R (The R Project for Statistical Computing, version 3.4.3), which was used for all remaining data processing.

## Carbon flux data processing

Soil respiration fluxes, in  $\mu$ mol CO $_2$  m $^{-2}$  s $^{-1}$ , were estimated using locally estimated scatterplot smoothing (LOESS) fit across observations as a function of lawn and mulch landcovers. For mulch and lawn landcover, observations from 2018 and the singular measurement from January 2019 were used. To account for uncertainties associated with human tampering and instrument stabilization, 36 of the 50 dendrometer bands were ultimately deemed reliable to include in our estimates of tree growth rates across the Charles River Campus, although one additional tree was removed from the study due to lack of end-of-season data on diameter at breast height (DBH). Tree stem and branch aboveground biomass was calculated using a generalized allometric equation for mixed hardwood species aboveground biomass:

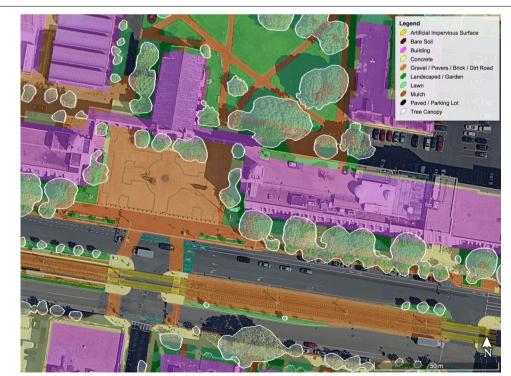


Fig. 1 Landcover type classification on the Marsh Plaza and BU Beach areas of Boston University's Charles River Campus

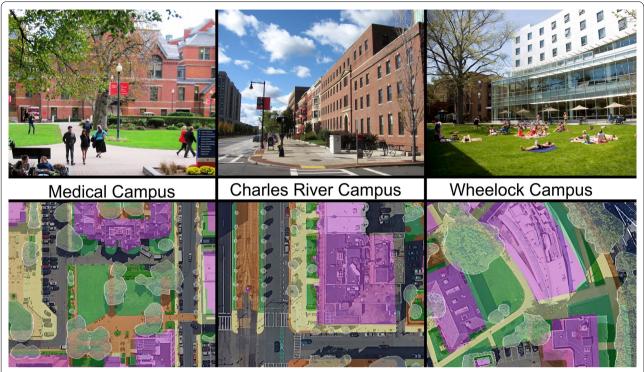


Fig. 2 Digitization examples from each campus; the Medical Campus (32 ha), the Charles River Campus (74 ha), and the Wheelock Campus (3.2 ha). Refer to Fig. 1 for landcover types and colors. (Image Credits L to R: BU Alzheimer's Disease Center, John Phelan, IEC Abroad)

$$bm = \exp(-2.4800 + 2.4835\ln dbh) \tag{1}$$

where *bm* is total aboveground biomass (in kg) and *dbh* equals the DBH (in cm) [29]. The component contribution of foliage:

$$foliage\ ratio = Exp(-4.0813 + 5.8816/dbh)$$
 (2)

was subtracted from total aboveground biomass [29]. For each of the 35 trees, the change in biomass (increment) was calculated by the difference between the initial and final biomass and divided by the total amount of time between measurements. The aboveground biomass increment over the 2018 growing season was multiplied by 0.5 to convert biomass to C mass [30]. Annual leaf production was estimated based on a locally observed ratio of woody growth to leaf production of  $0.50 \pm 0.01$  at the nearby Harvard Forest [31]; following Trlica et al. [32] we doubled the estimate of tree C uptake to account for leaf production. Accounting for foliar C as uptake is critical because C fixed in leaves is later released during the soil respiration process. Inclusion of respiration losses without calculating foliar C uptake would result in fundamentally biased estimates.

Rates of C sequestration by turf grasses across the Charles River Campus were calculated using sequestration estimates from Kaye et al. [33], who quantified C

sequestration on well-maintained urban lawns in Fort Collins, Colorado. Kaye et al. [33] determined that the C sequestered by mowed grass over the course of the growing season was  $133~{\rm g~C~m^{-2}~year^{-1}}$ , while unmowed, end-of-season grass sequestered 42 g C m $^{-2}$  year $^{-1}$ . These values were scaled to the Boston University urban campuses using the lawn area of each of the University's campuses. We are unaware of any analogous published studies closer to Massachusetts.

For each of the 35 trees equipped with dendrometer bands, a polygon was drawn in Google Earth to represent the tree canopy area to estimate the ground area (in m<sup>2</sup>) covered by each tree's canopy. Each tree's estimated annual C uptake (in kg year<sup>-1</sup>) was divided by its canopy area (in m<sup>2</sup>) to estimate C uptake as a function of tree canopy area across each of Boston University's urban campuses. C uptake was estimated on the basis of both the total land area and the amount of pervious land area using the digitized landcover. Monthly growth increments of C uptake were estimated by determining monthly growth rates for each tree across the 2018 growing season. These resulting growth increment calculations were used for both growth by aboveground woody biomass as well as grass growth on campus. Monthly tree growth increments were used to allow for the breakdown of annual woody biomass growth into discretized

months so monthly lawn and mulch respiration (C loss) and sequestration in trees and grass (C gain) could be compared. Additionally, utilizing in situ measurements of tree growth allowed for measurement of these monthly growth increments and increased local accuracy. Using Google Earth for polygon creation—including canopy extent—ensured that all open-source remote data were used, allowing Boston University community members to easily visualize the landcover and allowing other universities to conduct similar assessments of their landcover.

Net C exchange was calculated by subtracting the aboveground C sequestration (into woody biomass, leaves, and grass) for the year from the total soil respiration in units of Mg C year<sup>-1</sup>. This calculation was completed for each of the three urban campuses owned by Boston University. Positive values indicate net fluxes of C to the atmosphere, while negative values indicate net fluxes of C into vegetation.

# Anthropogenic CO<sub>2</sub> emissions calculations

We estimated both combined Scope 1 emissions from Boston University buildings (direct, local combustion) and Scope 2 emissions associated with electricity usage (emitted at a power plant elsewhere) for four of Boston University's campuses in 2016. Scope 1 emissions attributable to Boston University from 2016—the most recent available year of data-were independently estimated based on the Anthropogenic Carbon Emissions System (ACES), which calculated kg of CO<sub>2</sub> produced by all commercial and institutional buildings within a Census Block Group [34]. Estimates of kg  ${\rm CO}_2$  from the ACES model were divided by the total building floor area (m<sup>2</sup>) within each of the block groups, obtained from the Environmental Protection Agency's Hazus database [35], to produce the average kg  $\mathrm{CO}_2$  per square meter of building space per year for each census block. This average was then multiplied by building area within each Level 3 Assessors' parcel—obtained from MassGIS Data—on the Boston University Charles River, Medical, and Wheelock Campuses.

We also obtained a second, independent  $\mathrm{CO}_2$  emissions estimate for 2016 from Boston University's Sustainability Program, which tracks University-wide greenhouse gas emissions (Scope 1 and 2) as metric tons of carbon dioxide equivalents (MTCO $_2$ e) and breaks these emissions down into the Charles River Campus and the Medical Campus by the utility percentage. The University does not meter most individual buildings or even individual campuses. Emissions data for the Charles River Campus also include the Tanglewood and Wheelock Campuses, and the Medical Campus measurements are further broken down into the Medical Campus and the National Emerging Infectious Diseases Laboratories (NEIDL).

Because NEIDL is on the Medical Campus, these two estimates were grouped into one integrated estimate of the emissions from the Boston University Medical Campus. For both the ACES and University-based emissions, MgCO<sub>2</sub> and MgCO<sub>2</sub>e were converted into MgC using molecular weights. Both estimates contained annual emissions as this was the smallest temporal unit for which data was readily available.

#### Statistical analyses

All statistical analyses were conducted in R. All reported error values represent 95% confidence intervals. Spatially aggregated estimates were weighted based on the amount of associated landcover characteristics; errors were calculated as a root mean square to combine errors across multiple sources. Shapiro-Wilkes Test for normality showed that soil respiration data were not normally distributed for both mulch and lawn landcover (p=4.7 e-10, p=7.8 e-05), and that C sequestration rates in kg C m<sup>-2</sup> of canopy cover were also not normally distributed (p=3.1 e-05). Linear mixed effects models were used to conduct t-tests using Satterthwaite's method, providing p-values in the comparison of flux rates for mulched and lawn landcover, and for comparing flux rates between years. Within each year, landcover types were set as a fixed effect with day-of-year as a random effect, and between years both landcover type and year were set as fixed effects.

#### Results

Across the three urban Boston campuses, paved, building, and concrete surfaces are dominant across the landscape (Table 2). The remaining landcover was composed of more permeable materials such as lawn, landscaped garden, and brick or gravel (Fig. 3). The Charles River Campus has 22.4% pervious surface area and 12.2% canopy coverage. The Wheelock Campus is less than 1 km away, but resembles a more quintessential college campus as it has proportionally more lawn and mulched areas than the other urban campuses; Wheelock Campus has 29.8% pervious surface area and 31.6% canopy coverage. By contrast, the Medical Campus is the most highly developed campus, and it includes extensive hospital infrastructure and developed landcover with 20.5% pervious cover, and 8.9% canopy coverage.

Soil respiration rates in lawn areas averaged 5.5  $\pm$  0.5  $\mu$ mol CO $_2$  m $^{-2}$  s $^{-1}$  and 5.0  $\pm$  0.4  $\mu$ mol CO $_2$  m $^{-2}$  s $^{-1}$  for 2017 and 2018, respectively (Fig. 4). Respiration rates for mulch averaged 6.8  $\pm$  0.7 and 6.3  $\pm$  0.6  $\mu$ mol CO $_2$  m $^{-2}$  s $^{-1}$  for 2017 and 2018, respectively. Between 2017 and 2018, average flux rates were not significantly different (p=0.35). Within both 2017 and 2018, rates of respiration in mulched landcover were significantly higher

Table 2 Extent of lawn and mulched landcover on each of Boston University's urban campuses, as well as total biogenic emissions and sequestration, net biogenic C flux, and ACES-derived Scope 1 emissions

Campus (area)	Mulch cover (%)	Lawn cover (%)	Emissions via mulch (Mg C ha <sup>-1</sup> year <sup>-1</sup> )	Emissions via lawn (Mg C ha <sup>-1</sup> year <sup>-1</sup> )	C sequestered (Mg C ha <sup>-1</sup> year <sup>-1</sup> )	Net Biogenic C flux (Mg C ha <sup>-1</sup> year <sup>-1</sup> )	Scope 1 emissions (Mg C ha <sup>-1</sup> year <sup>-1</sup> )
Charles River Cam- pus (74 ha)	4.3	6.7	0.7 ± 0.18	0.9±0.25	0.9±0.67	0.8 ± 0.74	248
Medical Campus (32 ha)	4.8	10.1	$0.8 \pm 0.21$	$1.4 \pm 0.38$	$0.7 \pm 0.44$	$1.5 \pm 0.62$	440
Wheelock Campus (3.2 ha)	11.4	15.8	$1.9 \pm 0.49$	$2.2 \pm 0.60$	2.2±0.67	1.9 ± 1.02	201

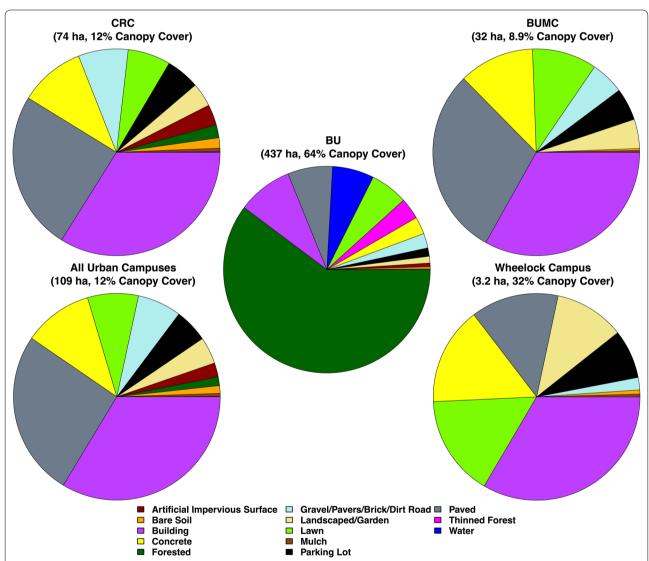
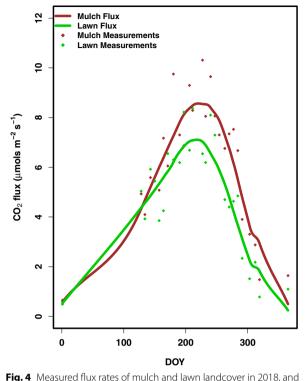


Fig. 3 Proportional landcover for each urban campus at Boston University, as well as an average of all urban campuses. Campus canopy cover and campus areas in hectares are also listed



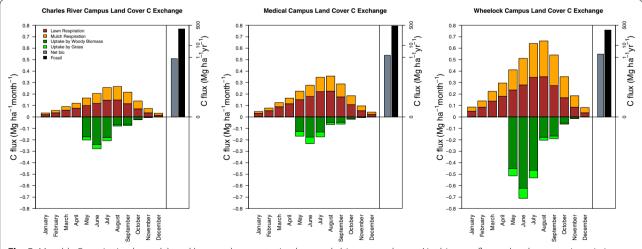
**Fig. 4** Measured flux rates of mulch and lawn landcover in 2018, and the LOESS curve used to predict flux rates year-round

than lawn landcover (p=0.012, p=0.00012, respectively). Rates of respiration were greatest in the months of July and August (Fig. 5).

Aggregating soil respiration fluxes from both land-covers, we find that the Charles River Campus emitted the least biogenic C per unit area, whereas the Wheelock Campus emitted the most biogenic C per unit area (Table 2). Further, each of the campuses contained more lawn than mulched landcover, though for the Charles River Campus and the Wheelock Campus, mulched and lawn landcover added roughly the same amount of C to the overall flux of  $CO_2$  to the atmosphere.

Woody biomass sequestered  $16.8 \pm 6.9$  kg C year<sup>-1</sup> tree<sup>-1</sup>, and C sequestration rates were not statistically significant between trees planted in mulched and lawn landcovers across 2017 and 2018 (p=0.50). C sequestered by trees was greatest in May, June, and July, and was modeled to be zero from December through April. On both the Charles River Campus and the Wheelock Campus, C loss through respiration was higher than sequestration each month of the year except for May and June, when C sequestration into woody biomass and grass outweighed C loss (Fig. 5). On the Medical Campus, C loss through respiration was higher than sequestration during each month of the year. Consequently, each campus was an annual net source of biogenic C.

Using the ACES model, we found that Scope 1 building heating related emissions on the Charles River Campus were 248 Mg C ha $^{-1}$  year $^{-1}$  in 2016, the Medical Campus was 440 Mg C ha $^{-1}$  year $^{-1}$ , and the Wheelock Campus emitted 201 Mg C ha $^{-1}$  year $^{-1}$ . University electricity usage data – which combines the Charles



**Fig. 5** Monthly C respiration by mulch and lawn and sequestration by woody biomass and grass. Net biomass flux and anthropogenic emissions estimates are represented on a logarithmic scale. All numbers are per hectare, as they have been normalized by the unit area of the respective campus

River, Tanglewood, and Wheelock Campuses – showed that the three campuses produced a combined 263 Mg C  $ha^{-1}$  year<sup>-1</sup> and the Medical Campus emitted 258 Mg C  $ha^{-1}$  year<sup>-1</sup> in 2016.

Additionally, there was significant variation in the proportion of emissions through biogenic respiration to anthropogenic emissions. On the Charles River Campus, respiration equaled 0.67% of Scope 1 emissions, whereas respiration represented 0.51% of Scope 1 emissions on the Medical Campus and 2.06% of Scope 1 emissions on the Wheelock Campus (Table 2). This calculation was repeated to find the proportion of net biogenic emissions to anthropogenic emissions.

#### Discussion

Climate action plans allow cities and institutions to create an achievable framework for reducing their greenhouse gas emissions. Specifically, these plans provide targets for emissions reductions within a given timeframe and implementation pathways for reducing emissions through clean technologies, renewable energies, changes in transportation patterns, and the use of green infrastructure. Boston University owns roughly 0.76% of the combined land areas of the City of Boston and the bordering Town of Brookline, a sizeable portion of metropolitan Boston's land for a single land owner, making it particularly important that the University accurately accounts for C fluxes since these fluxes can have a significant impact on the City of Boston as a whole. Our estimates across the three Boston University campuses show that all were a net source of biogenic C to the atmosphere due to landscaping management practices and that biogenic emissions comprised 0.51% to 2.06% of the total anthropogenic emissions on each campus. This is a lower-bounds estimate as it does not include direct emissions associated with maintenance such as lawn mowing, transportation, and leaf blowing. Considering these sources, lawn mowing and fertilizer application alone increase Charles River Campus biogenic C emissions by 1.0% under typical non-irrigated "home" lawn management practices [18], with an additional 2.3 to 9.5 kg C in maintenance for each tree each year [19, 36].

Lawns comprise a higher proportion of landcover than mulch on each of the three urban campuses, although on the Charles River Campus and Wheelock Campus overall C loss was roughly equal between mulch and lawn due to the higher rate of C flux per unit area from mulched areas. The Wheelock Campus' high rate of C respiration was attributed to its high overall pervious landcover fraction. Nearly a third of the Wheelock Campus is covered by lawn, mulch, or soil, whereas the Charles River Campus and the Medical Campus had lower overall rates of C respiration due to the prevalence of impervious surfaces

(77.6% and 79.5%, respectively). However, Wheelock also had over twice the amount of canopy cover as the Charles River Campus and nearly four times that of the Medical Campus, offsetting 47.3% of respiratory emissions of  $CO_2$ .

Both the Charles River Campus and the Medical Campus had substantially lower tree canopy cover (12.2% and 8.9%, respectively) than the City of Boston's average of 26% [3]. Sequestration into woody biomass and grass offset 54.0%, 52.7%, and 32.7% of biogenic emissions from soil respiration for the Wheelock Campus, Charles River Campus, and Medical Campus, respectively.

Of Boston University's urban campuses, Wheelock Campus most closely resembles a "traditional" college campus with a large, manicured grass quad surrounded by campus buildings (Fig. 2). The relatively high proportion of vegetation and canopy cover facilitates landcover modification. Over half of Wheelock Campus's biogenic C emissions are offset by on-campus tree and grass sequestration, suggesting that smaller changes in oncampus landcover management strategies such as minimizing mulch application and increasing tree canopy cover could contribute to Wheelock achieving biogenic C neutrality or even the potential creation of a small biogenic sink.

Although transportation of goods, services, and University community members to and from campus represent a significant source of University-based anthropogenic emissions, granular emissions data from fossil fuel combustion were not available for this analysis. The Boston University Climate Action Plan focuses primarily on Scope 1 and Scope 2 emissions, and estimates that University-based indirect personal transportation, purchasing, and waste disposal emissions are approximately 200,000 MTCO<sub>2</sub>e (54,582 Mg C) [12]. Including Scope 1 and Scope 2 emissions, over which the University maintains greater control, allows for the most accurate estimates of University-related anthropogenic emissions and represents the emissions most easily regulated by the University.

The ACES emissions inventory is at a census block scale, including areas beyond the campus boundary. The University's own emissions inventory aggregates across campuses since it is not sub-metered and includes both direct Scope 1 building emissions and indirect Scope 2 electricity emissions. Many large, institutional land owners often do not have sub-meters for their many buildings, making it difficult to disaggregate emissions data and assess efficiency improvements. The ACES model estimates for Scope 1 C emissions in the Medical Campus were significantly higher than the Charles River and Wheelock Campuses, likely due to the high proportion of industrial buildings—which were modeled with higher

 $\mathrm{CO}_2$  emission rates than residential buildings—in the Medical Campus' census block. While the exact emissions are more uncertain than would be desirable, it is clear that Scope 1 and Scope 2 emissions far surpass biogenic emissions on each urban campus. For example, biogenic emissions from lawn and mulched landcover on the Medical Campus represent just 0.51% of Scope 1 C emissions from the Medical Campus. On the Wheelock Campus, which has considerably lower anthropogenic C emissions, biogenic emissions make up 2.06% of all C emissions.

Although a breakdown of anthropogenic emissions into smaller temporal/spatial intervals is unavailable in this study, seasonal variability is incorporated into emissions estimates from ACES and the University-based emissions inventory. ACES, for example, found that summertime anthropogenic emissions represent 57% of wintertime emissions, largely due to the lack of heating-related emissions during the summer months [6, 37]. Scheduled variations in on-campus activity—such as holiday breaks, graduation, and move-in weekend-also likely affect campus anthropogenic emissions, although the campus is heavily populated year-round, including during the intersession summer months. While these variations are not explicitly included in ACES emissions estimates, they are accounted for in the University's emissions inventory. Variations in landscape management also occur sporadically over the course of the growing season, particularly during campus "beautification" initiatives preceding high-profile campus events such as parents' weekend, alumni weekend, and commencement weeks. This suite of management and seasonal variations is included in our measurements of biological fluxes of soil respiration and tree growth across an entire season of landscape management. It is also worth noting that management intensity may slightly vary between campuses due to the relative prevalence of high-profile campus events on some campuses.

Considering C sinks on non-urban lands – such as Boston University's Sargent Center and Tanglewood Campuses—affects Boston University's entire C budget. The Sargent Center and Tanglewood Campuses, for example, include 2.76 km² of forest or thinned forest, which helps to offset 4,700 Mg CO<sub>2</sub> (1,283 Mg C) of the University's C emissions [12]. This study focused on Boston University's three urban campuses, but a University-wide assessment should include C respiration and sequestration estimates from non-urban satellite campuses to quantify University C fluxes across all lands owned by universities. The question of additionality—or the allowance of C credits for business-as-usual management of forests—needs to be considered in both urban and non-urban satellite campuses [38]. This approach might allow universities such

as Boston University to claim C credits for maintaining forested land as a C sink, offsetting biogenic or anthropogenic emissions from their urban campuses. However, this C credit can only be issued under the recognition that the University maintained this sink for the purpose of C sequestration, as opposed to receiving credit for simply following a "business-as-usual" plan.

This analysis points to several important data shortfalls and opportunities for managing C fluxes. Mulched soils have a significantly higher rate of C loss from soil respiration than lawns, indicating that replacing mulch with ground cover plants or even lawn-minimizing the application of mulch on campuses—has the potential to greatly reduce C emissions from soil respiration. Preceding high-profile campus events, mulch is very liberally applied across the campus for "beautification." Applying a thinner layer of mulch or applying biennially will assist with reducing soil respiration rates. Although C respiration rates vary based on the amount of mulch applied as well as other maintenance methods and weather patterns-this study included growth and respiration measurements over multiple years with varying temporal frequencies to capture the timing and controls on biogenic fluxes. The observed inter-annual variability in measured C fluxes is within the range typically observed in natural systems [39].

We do note that mulch does have some environmental benefits that we have not accounted for here, particularly insulating soils to retain moisture [40]. However, most of the Boston University campus is well irrigated with timed sprinklers, reducing the need for hydrological benefits from mulch. Planting additional trees has potential to mitigate some of the high losses of  $\mathrm{CO}_2$  from soil respiration on University campuses. Two of the urban Boston University campuses (i.e. the Charles River and Medical Campuses) have particularly low canopy cover compared to the City of Boston; there are opportunities to increase canopy coverage to help offset on-campus emissions and potentially reduce building energy demand through canopy shading effects [13–15, 41].

Critical components not included in this study include investigating the tradeoffs between increased canopy cover and respiration rates, as some potential benefits of increasing canopy cover could be offset by increased stem and root respiration. While we did not undertake a full life cycle assessment of the C fluxes associated with mulch or the C emissions associated with lawn or tree maintenance [18, 19], the Scope 1 implications are clear nonetheless. Past studies [18, 42, 43] show that both lawn mowing and fertilizer use represent hidden C costs that reduce the overall C sequestration capacity of lawns. Generally, lawns are net C negative (resulting in additional C release to the atmosphere), though reducing

C-intensive landscaping techniques would facilitate greater C sequestration potential.

While large-scale reductions of Boston University's fossil fuel emissions are central to achieving net C neutrality, biogenic fluxes represent an important component of the greenhouse gas budget. Land owners such as universities should consider their biogenic emissions and the role of their management choices in both storing C and creating large Scope 1 sources through import and release of C through mulch, and the potential for uptake through canopy expansion.

As global climate policies slowly advance, it has become increasingly important for independent institutions to contribute to reductions in C emissions to achieve the established goal of limiting global temperature rise to 2 °C or less above pre-industrial levels [44]. Although Boston University has pledged to reduce fossil fuel emissions through a series of offsets and renewable energies, the institution should ensure that local C fluxes are also accounted for in its University-wide C quantification. While biogenic C fluxes represent a small proportion of the total C emissions from Boston University, its urban campuses cannot achieve true net C neutrality without first considering biogenic C fluxes. Further, net biogenic C neutrality is more readily feasible if the University adopts revised approaches to beautification of the campus through landcover management.

#### **Conclusions**

Green infrastructure and nature-based solutions cannot be assumed to result in biogenic C sinks. We show that C fluxes from mulched and lawn landcover via soil respiration on Boston University's three urban campuses outweigh C sequestration by on-campus vegetation, making each of the University's urban campuses net biogenic C sources. Emissions via soil respiration were found to be significantly higher in mulch than lawn landcover per unit area, although on two campuses, lawn and mulch contributed roughly equal amounts of C to the overall CO<sub>2</sub> flux to the atmosphere due to a larger proportion of lawn landcover. Additionally, we show that although anthropogenic emissions are considerably larger than biogenic emissions, biogenic fluxes into grass and woody biomass still have the potential to offset a portion of anthropogenic emissions while providing additional ecosystem services. Finally, we conclude that biogenic C emissions from urban university campuses can be reduced through reduction in the frequency and quantity of mulch addition onto urban campuses and through increasing the number of trees.

# **Supplementary information**

**Supplementary information** accompanies this paper at https://doi. org/10.1186/s13021-020-00160-5.

Additional file 1. Landcover on Boston University's urban campuses in the City of Boston overlaid on high-resolution aerial imagery from MassGIS (https://docs.digital.mass.gov/dataset/massgis-data-usgs-color-ortho-imagery-2019). The tree canopy layer has been made transparent to allow for visualization of layered landcovers when zoomed in.

#### **Abbreviations**

C: Carbon; CO<sub>2</sub>: Carbon dioxide; PVC: Polyvinyl chloride; LOESS: Locally estimated scatterplot smoothing; LiDAR: Light detection and ranging; *bm*: Biomass; DBH: Diameter at breast height; ACES: Anthropogenic carbon emissions system; MTCO<sub>2</sub>e: Metric tons of carbon dioxide equivalent; NEIDL: National emerging infectious diseases laboratories.

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#### Authors' contributions

WJH discretized and digitized the Boston University Campuses and was the primary author. LRH and PHT conceived the research; VC, AHYB, ML, and IAS installed measurement tools and sensors and collected measurements. CKG compiled and analyzed anthropogenic emissions data. WJH, LRH, PHT, IAS, and ML designed and carried out the analysis and data modeling. All authors read and approved the final manuscript.

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## Availability of data and materials

The data is published on the Harvard Dataverse at https://dataverse.harvard.edu/dataset.xhtml?persistentId=doi%3A10.7910%2FDVN%2FRUSGG0.

#### **Competing interests**

The authors declare no conflicts of interest. Funding sponsors played no part in designing the study, collecting, analyzing or interpreting data, and in publishing the results.

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## References

- 1. We Are Still In. https://www.wearestillin.com/. Accessed 26 Jan 2020.
- Hutyra LR, Duren R, Gurney KR, Grimm N, Kort EA, Larson E, et al. Urbanization and the carbon cycle: current capabilities and research outlook from the natural sciences perspective. Earths Future. 2014;2:473–95.

- Raciti SM, Hutyra LR, Newell JD. Mapping carbon storage in urban trees with multi-source remote sensing data: relationships between biomass, land use, and demographics in Boston neighborhoods. Sci Total Environ. 2014;500–501;72–83.
- Hardiman BS, Wang JA, Hutyra LR, Gately CK, Getson JM, Friedl MA. Accounting for urban biogenic fluxes in regional carbon budgets. Sci Total Environ. 2017;592:366–72.
- 5. Bergeron O, Strachan IB.  $\rm CO_2$  sources and sinks in urban and suburban areas of a northern mid-latitude city. Atmos Environ. 2011;45:1564–73.
- Decina SM, Hutyra LR, Gately CK, Getson JM, Reinmann AB, Short Gianotti AG, et al. Soil respiration contributes substantially to urban carbon fluxes in the greater Boston area. Environ Pollut. 2016;212:433–9.
- McRae JE, Graedel TE. Carbon dioxide in the urban atmosphere: dependencies and trends. J Geophys Res Oceans. 1979;84(C8):5011–7.
- Sargent M, Barrera T, Nehrkorn T, Hutyra LR, Gately CK, Jones T, et al. Anthropogenic and biogenic CO<sub>2</sub> in the Boston urban region. Proc Natl Acad Sci. 2018:115(29):7491–6.
- Deetjen TA, Conger JP, Leibowicz BD, Webber ME. Review of climate action plans in 29 major US cities: comparing current policies to research recommendations. Sust Cities Soc. 2018;41:711–27.
- Center for Climate and Energy Solutions. Mayors Leading the Way on Climate: How Cities Large and Small are Taking Action. 2018. www.c2es.
- Chen WY. The role of urban green infrastructure in offsetting carbon emissions in 35 major Chinese cities: a nationwide estimate. Cities. 2015;44:112–20
- Recommendations of the Climate Action Plan Task Force for Boston University's Climate Action Plan. 2017. https://www.bu.edu/climateactionplan/climate-action-plan-task-force/report/.
- Akbari H. Shade trees reduce building energy use and CO<sub>2</sub> emissions from power plants. Environ Pollut. 2016;116:S119–26.
- Nowak DJ, Appleton N, Ellis A, Greenfield E. Residential building energy conservation and avoided power plant emissions by urban and community trees in the United States. Urban For Urban Green. 2017;21:158–65.
- McPherson EG, Rowntree RA. Energy conservation potential of urban tree planting. J Arbor. 1993;19(6):321–31.
- Churkina G. The role of urbanization in the global carbon cycle. Front Ecol Evol. 2016;3(144):1–9.
- Janssens IA, Lankreijer H, Matteucci G, Kowalski AS, Buchmann N, Epron D, et al. Productivity overshadows temperature in determining soil and ecosystem respiration across European forests. Glob Chang Biol. 2001;7:269–78.
- Selhorst A, Lal R. Net carbon sequestration potential and emissions in home lawn turfgrasses of the United States. Environ Manage. 2013;51:198–208.
- Nowak DJ, Stevens JC, Sisinni SM, Luley CJ. Effects of urban tree management and species selection on atmospheric carbon dioxide. J Arboric. 2002;28(3):113–22.
- Briber BM, Hutyra LR, Reinmann AB, Raciti SM, Dearborn VK, Holden CE, et al. Tree productivity enhanced with conversion from forest to urban land covers. PLoS ONE. 2015;10(8):e0136237.
- 21. Smith IA, Dearborn VK, Hutyra LR. Live fast, die young: accelerated growth, mortality, and turnover in street trees. PLoS ONE. 2019;14(5):e0215846.
- Gough CM, Elliott HL. Lawn soil carbon storage in abandoned residential properties: an examination of ecosystem structure and function following partial human-natural decoupling. J Environ Manage. 2012;98:155–62.
- 23. Jo H-K. Impacts of urban greenspace on offsetting carbon emissions for middle Korea. J Environ Manage. 2002;64(2):115–26.
- Alonzo M, McFadden JP, Nowak DJ, Roberts DA. Mapping urban forest structure and function using hyperspectral imagery and lidar data. Urban For Urban Green. 2016;17:135–47.

- Wilkes P, Disney M, Vicari MB, Calders K, Burt A. Estimating urban above ground biomass with multi-scale LiDAR. Carbon Balance Manag. 2018:13:10.
- Zhou W, Fisher B, Pickett STA. Cities are hungry for actionable ecological knowledge. Front Ecol Environ. 2019;17(3):135.
- United States Census Bureau. Patterns of Metropolitan and Micropolitan Population Change: 2000 to 2010 Tables. 2017. https://www.census.gov/ data/tables/time-series/dec/c2010sr-01.html. Accessed 20 Oct 2020.
- Data Tools: 1981–2010 Normals. National Climatic Data Center. https:// www.ncdc.noaa.gov/cdo-web/datatools/normals. Accessed 1 Feb 2019.
- Jenkins JC, Chojnacky DC, Heath LS, Birdsey RA. Comprehensive Database of diameter-based biomass regressions for North American Tree Species. Gen Tech Rep. 2004;319:45.
- 30. Chave J, Coomes D, Jansen S, Lewis SL, Swenson NG, Zanne AE. Towards a worldwide wood economics spectrum. Ecol Lett. 2009;12:351–66.
- Munger W, Wofsy S. Biomass Inventories at Harvard Forest EMS Tower since 1993. 2019. Harvard Forest Data Archive: HF069. https://harvardforest1.fas.harvard.edu/exist/apps/datasets/showData.html?id=hf069.
- Trlica A, Hutyra LR, Morreale LL, Smith IA, Reinmann AB. Current and future biomass carbon uptake in Boston's urban forest. Sci Total Environ. 2020:709:136–96
- Kaye JP, McCulley RL, Burke IC. Carbon fluxes, nitrogen cycling, and soil microbial communities in adjacent urban, native and agricultural ecosystems. Glob Chang Biol. 2005;11(4):575–87.
- Gately CK, Hutyra LR. CMS: CO<sub>2</sub> emissions from fossil fuels combustion, ACES inventory for Northeastern USA. 2018. Ornl DAAC. https://doi. org/10.3334/ORNI DAAC/1501.
- HAZUS Database. Federal Emergency Management Agency, Washington, DC. 2018. https://www.fema.gov/flood-maps/tools-resources/flood-map-products/hazus. Accessed 1 Sept 2018.
- McPherson EG, Kendall A. A life cycle carbon dioxide inventory of the Million Trees Los Angeles program. Int J Life Cycle Ass. 2014;19:1653–65.
- Gately CK, Hutyra LR. Large uncertainties in urban-scale carbon emissions. J Geophys Res Atmos. 2017;122:11242–60.
- 38. Powers J, Haddon A. The role of RECs and additionality in green power markets. Schneider Electric. 2018;22:65.
- Finzi AC, Giasson MA, Barker-Plotkin AA, Aber JD, Boose ER, Davidson EA, et al. Carbon budget of the Harvard Forest Long-Term Ecological Research sit: pattern, process, and response to global change. Ecol Monogr. 2020. https://doi.org/10.1002/ecm.1423.
- Iles JK, Dosmann MS. Effect of organic and mineral mulches on soil properties and growth of Fairview flame red maple trees. J Arboric. 1999;25(3):163–7.
- Spronken-Smith RA, Oke TR. The thermal regime of urban parks in two cities with different summer climates. Int J Remote Sens. 1998;19(11):2085–104.
- 42. Zirkle G, Lal R, Augustin B. Modeling Carbon Sequestration in Home Lawns. HortScience. 2011;46(5):808–14.
- Strohbach MW, Arnold E, Haase D. The carbon footprint of urban green space—a life cycle approach. Landscape Urban Planning. 2012;104:220–9.
- 44. United Nations. Paris Agreement. 2015. https://unfccc.int/process-and-meetings/the-paris-agreement/the-paris-agreement.

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