



## Research Article

# Enhanced photocatalytic reduction of toxic Cr(VI) with Cu modified ZnO nanoparticles in presence of EDTA under UV illumination

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## Abstract

In this study, the photocatalytic reduction of Cr(VI) to Cr(III) in aqueous solution using 1 wt% Cu/ZnO photocatalyst in presence of ethylenediaminetetraacetic acid (EDTA) under UV irradiation was investigated. The photocatalytic Cr(VI) reduction with ZnO under UV irradiation was significantly low. A fast Cr(VI) reduction with ZnO was recorded in presence of EDTA. Results demonstrated that the rate constant raised from  $1.6 \times 10^{-2}$  to  $7.5 \times 10^{-2} \text{ min}^{-1}$  with 20 ppm EDTA addition. The optimum reduction (90%) was observed within 30 min at pH 3. The reduction efficiency of Cr(VI) was evidently increased from 6% to 99%, with increasing ZnO dosages up to 40 mg. The role of co-catalysts such as Au, Ag, Cu and NiO was also studied. The addition of Cu ion remarkably accelerated the reduction of Cr(VI), compared with ZnO/EDTA system alone. The Cu motivated ZnO/EDTA system completed the 99% reduction within only 30 min. The Cu ions were reduced to metallic copper and  $\text{Cu}_2\text{O}$  by conduction band electrons of ZnO. The electron-hole recombination was gently declined by the shallow electron trapping of Cu. On the other hand, EDTA could be performed in two ways by forming Cr(III)-complex and separating Cr(III) from ZnO surface and by consuming valence band holes. The Cu modified ZnO photocatalyst was characterized by  $\text{N}_2$  adsorption-desorption isotherm, TEM, EDX, XPS, DRS and PL spectra. We evaluated the influence of pH, photocatalyst dosage, initial concentration of EDTA and initial concentration of  $\text{Cu}^{2+}$  ion (Cu doping amount) on photocatalytic reduction of Cr(VI). The pseudo first order kinetic reaction was confirmed for this reduction process. Herein, a probable mechanism of this photocatalytic reduction process was also presented.

**Keywords** Photocatalytic reduction · ZnO nanoparticles · EDTA · Electron-hole recombination · Cu modification

## 1 Introduction

The 21st most abundant element of the earth's crust is chromium. Although chromium naturally occur as metal, it is conventionally admitted as a toxic element. This notorious chromium level has been increased in surface water with the development of urbanization and

industrialization. Lather tanning, electroplating, metal corrosion inhibition, dye, pigment and alloy manufacturing industries release huge amount of chromium into nature every day [1, 2]. Between the two common oxidized forms such as hexavalent chromium (Cr(VI)) and trivalent chromium (Cr(III)), the Cr(III) is less pernicious than the Cr(VI) [3]. The non-biodegradability and high mobility of Cr(VI)

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in nature can present a mutagenic and carcinogenic risk for all living organisms [4]. On contrast, a trace amount of Cr(III) is required for human and animal health, since it is essential for sugar and lipid metabolism [5]. The World Health organization (WHO) firmly recommended that the maximum permissible level of Cr(VI) should be less than  $50 \mu\text{g/L}^{-1}$  in drinking water [6]. Hence, the removal of toxic Cr(VI) from wastewater has been necessitated immensely. Electrocoagulation, microfiltration, adsorption, ion exchange, chemical precipitation, and etc. are generally applied to eliminate Cr(VI) from industrial waste water [7–11]. Nevertheless, these multifarious processes generally require a large instrumentation and huge amount of chemicals and organic additives, all of which markedly interrupt their practical use. Moreover, the production of secondary pollutants is another drawback of these traditional methods [12, 13]. Amid the most useful strategies, photocatalytic reduction using semiconductor-based photocatalyst overcomes these mentioned limitations impressively. Photocatalytic reduction technology manifests a high reduction efficiency and low energy consumption. The low cost and simple application make this process quite attractive to reduce hazardous Cr(VI) to benevolent Cr(III) [14]. Furthermore, photocatalysts can be recycled and reused repeatedly after completing photocatalytic reduction process [15].

Numerous semiconductor photocatalysts, for instance,  $\text{TiO}_2$ , ZnO, CdS and  $\text{WO}_3$  have received great consideration for photocatalytic treatment of heavy metals, textile dyes and other toxic pollutants [16–19]. This photocatalytic method generally involves the migration of valence electrons to conduction band, when photocatalyst are irradiated under light having photons with more energy than bandgap of the photocatalyst. Migrated conduction band electrons can reduce the toxic heavy metal if the redox potential of reactant is more positive than the conduction band of photocatalyst [20]. ZnO is also an appealing semiconductor photocatalyst along  $\text{TiO}_2$ , since both have almost the same band gap ( $\sim 3.20 \text{ eV}$ ) and chemically stable properties [21, 22]. According to some researchers, ZnO shows better photocatalytic activity than  $\text{TiO}_2$  [23, 24]. The study demonstrated that ZnO nanoparticles can photoreduce toxic Cr(VI) efficiently to less toxic Cr(III). However, the application of ZnO powder during photoreduction of Cr(VI) still faces some limitations such as solubility below pH 3, poor dispersity and rapid combination of photoinduced electron–hole as well as narrow range of light response. Hence, it is essential to improve the photocatalytic activity of ZnO. The addition of sacrificial electron donors [25], composite formation such as ZnO/ $\text{TiO}_2$ , ZnO/CuS and ZnO/ $\text{ZnFe}_2\text{O}_4$  [26, 27] or/and metal doping can modify the photocatalytic properties of ZnO [28, 29]. Recently, some common sacrificial electron

donors, namely citric acid, EDTA, oxalic acid, formic acid and tartaric acid, have been tested on the reduction of Cr(VI) [3, 30]. Bao et al. explained it that EDTA gave the most positive effect on improving the Cr(VI) reduction compared with other sacrificial electron donors. Moreover, the number of carboxyl groups also controlled the reduction efficiency. Accordingly, EDTA enhanced Cr(VI) reduction more than the case with oxalate because oxalate readily oxidizes to  $\text{CO}_2$  during the photocatalytic process, while EDTA produces several intermediates in route to its mineralization, which may also contribute to the synergistic effect [31]. In addition, Westin and Rasmuson reported that the presence of ethylenediaminetetraacetic acid (EDTA) can control the sizes and morphologies of the particles owing to the chelating effects of EDTA with metal ions of solution [32]. Therefore, Cr(VI)/EDTA mixed system has been grabbed immense attention to accelerate the heterogeneous photocatalytic Cr(VI) reduction [33].

Among different metallic candidates, copper (Cu) is a very promising material for photocatalysts modification. Copper is an inexpensive material comparing with other noble metals (gold, platinum and silver) recognized as very efficient co-catalysts of ZnO [34]. Furthermore, copper shows similar properties as gold and silver, due to occupying in the same group of the periodic table and its electronic configuration and the face centered cubic (FCC) structure of the atom's location. Under irradiation, copper ion forms complexes with organic acids and ultimately reduces to copper metal [35]. Moreover, Cu doping into ZnO can modify its electrical, optical and photocatalytic activity significantly [28]. Therefore, copper can be applied as potential co-catalyst with ZnO nanoparticles to enhance the photocatalytic behavior of ZnO nanoparticles. Even though the photocatalytic activities of ZnO nanoparticles have been examined for degradation various organic pollutants as well as heavy metals, to the best of our knowledge, the performance of Cu motivated ZnO semiconductor for Cr(VI) photo-reduction in presence of EDTA under UV illumination has not been investigated yet.

Therefore, the paper work aims to go a step forward in the perception of the key step to maximize the photocatalytic reduction of Cr(VI) with Cu motivated ZnO photocatalyst in presence of EDTA under UV irradiation.

## 2 Experimental

### 2.1 Chemicals and materials

ZnO nanopowder (surface area  $15\text{--}25 \text{ m}^2\text{g}^{-1}$ , particle size  $50\text{--}70 \text{ nm}$ ) were purchased from SIGMA–ALDRICH Co. USA. One thousand ppm chromium ( $\text{Cr}^{6+}$ ), 1000 ppm copper ( $\text{Cu}^{2+}$ ) standard solution and 1,5-diphenylcarbazid

( $C_{13}H_{14}N_4O$ ) powder were brought from Wako Pure Chemicals Industries Ltd., Osaka, Japan. EDTA-2Na and acetone ( $C_3H_6O$ ) were obtained from Kishida Chemical Co. Ltd., Osaka, Japan. All chemical were used without further purification. NaOH (1.0 mol/L) and  $HNO_3$  (1.0 mol/L) were used to adjust solution pH.

## 2.2 Photocatalytic run

Photocatalytic reduction of Cr(VI) was conducted in a 50 mL Pyrex vessel as reactor placing on magnetic stirrer. A black light (Toshiba Lighting & Technology Corp., 15 W) was positioned on the side of reactor as light source with a maximum emission of about 352 nm. The light intensity was determined by a digital UV intensity meter (USHIO, UIT-201) with a sensor (UV-365PD, 330~390 nm), and the value was  $0.25 \text{ mW/cm}^2$ . For this treatment, 40 mg ZnO photocatalyst was dispersed into 30 mL aqueous solution of hexavalent chromium (10 mg/L). Then, 40 ppm EDTA as a hole scavenger and  $40 \mu\text{L}$   $Cu^{2+}$  ion (1000 ppm) as co-catalyst were added into solution at room temperature. In dark, the sample was continuously stirred on magnetic stirrer for 30 min to achieve adsorption-desorption equilibrium of Cr(VI) on the surface of Cu modified ZnO. At specific interval, 1.5 mL aliquot was taken out and centrifuged at 1000 rpm for 15 min instantly to suppress nano powder from the solution. The concentration of Cr(VI) was determined by the Neslerization method using UV-VIS spectrophotometer at 540 nm.

## 2.3 Characterization

The specific surface area was determined by the Brunauer-Emmett-Teller  $N_2$  adsorption-desorption isotherm measurement using BELSORP-mini II (BEL Japan). The morphology of ZnO and Cu/ZnO were observed by a transmission electron microscope (TEM; HITACHI S-4000, Japan). The elemental analysis of Cu/ZnO nanoparticles

has been prepared by energy dispersive X-ray spectroscopy (EDS). X-ray photoelectron Spectroscopy (XPS) measurement of Cu/ZnO photocatalyst was conducted on PHI Quantera SXM. To obtain band gap of photocatalyst, UV-VIS diffuse reflectance spectra (DRS) were recorded in the range of 200–870 nm on a UV spectrophotometer (V-750 SHIMAZU, Japan) equipped with an integrating sphere attachment, using  $BaSO_4$  as reference. Photoluminescence spectra (PL) were recorded on a Fluorescence Spectrophotometer (RF-5300PC, SHIMAZU).

For the reusability of photocatalysts, 1 wt% Cu/ZnO was collected after each cycle by centrifugation. Then were washed with deionized water and ethanol. Then dried catalyst 1 wt% Cu/ZnO was used for next run. The further procedure for Cr(VI) reduction process was the same as mentioned above, except for the addition of Cu ion.

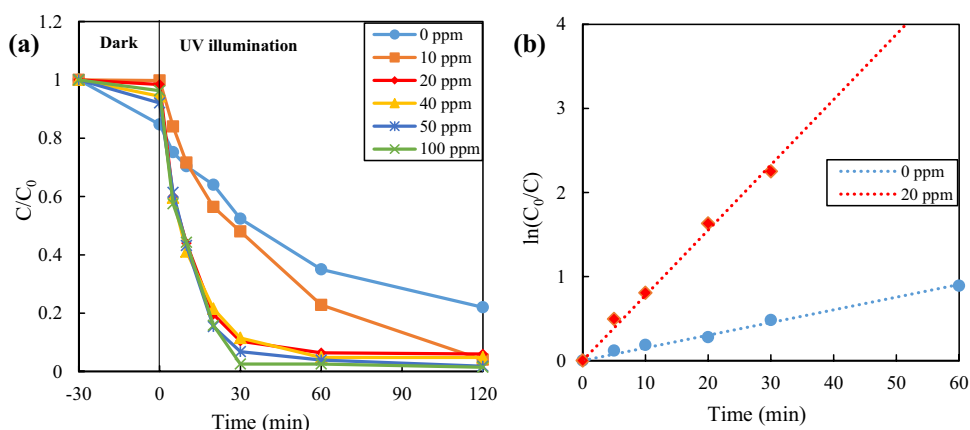
## 3 Result and discussion

### 3.1 Photocatalytic reduction of aqueous Cr(VI)

#### 3.1.1 Influence of EDTA addition

The influence of EDTA addition on Cr(VI) removal was investigated using 0 ppm to 100 ppm EDTA with 10 ppm Cr(VI) as initial concentration. Result was shown in Fig. 1. It explained that 78% photocatalytic reduction of Cr(VI) was achieved with ZnO after 120 min in the absence of EDTA, while this percentage raised to 98.5 after adding 10 ppm to 100 ppm EDTA. The improvement may be due to the fact that organic additives containing EDTA become an effective electron donor, which captured positively charged holes ( $h^+$ ) generated in valence band and more negatively charged electrons participated in the reduction of Cr(VI) simultaneously [36]. Consequently, reduction of Cr(VI) was accelerated. On contrary, Litter et al. [31] explained that the Cr(III) deposition onto photocatalyst

**Fig. 1** Effect of EDTA addition on the photocatalytic reduction of Cr(VI) with ZnO (a); The plot of  $\ln(C_0/C)$  versus irradiation time (b)



surface after reduction was inhibited due to the formation of Cr(III)-EDTA complex for the advanced photocatalytic activity of photocatalyst. Since the reduction rate of Cr(VI) over 20 to 100 ppm EDTA was almost the same, more than 20 ppm of EDTA was favorable for this experiment.

The photocatalytic reaction kinetics has been elucidated by Langmuir–Hinshewood (L–H) model explaining the relationship between initial concentration of Cr(VI) and reduction rate of Cr(VI). If the Cr(VI) concentration is very low, that is,  $1 \gg K[\text{Cr(VI)}]$ , the reduction rate will aid to pseudo-first-order kinetic law [25, 37].

$$r = \frac{d[\text{Cr(VI)}]}{dt} = kK[\text{Cr(VI)}] = k_{\text{obs}}[\text{Cr(VI)}] \tag{1}$$

$$-\ln \frac{[\text{Cr(VI)}]}{[\text{Cr(VI)}]_0} = k_{\text{obs}}t$$

where,  $[\text{Cr(VI)}]_0$  is the initial Cr(VI) concentration,  $t$  is the irradiation time and  $k_{\text{obs}}$  is the pseudo-first-order rate constant ( $\text{min}^{-1}$ ).

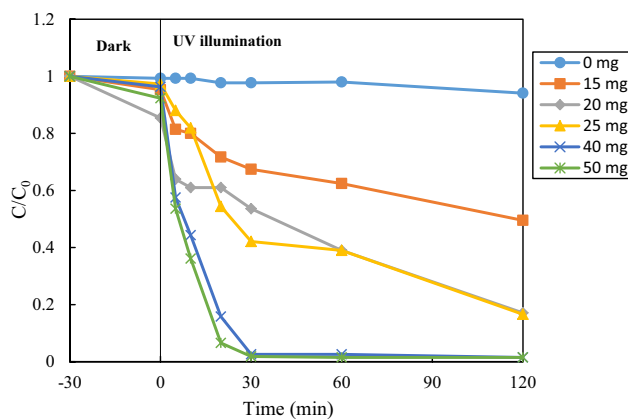
The kinetics of this photocatalytic reduction of Cr(VI) with ZnO in presence of 20 ppm of EDTA was explained in Fig. 1b where,  $\ln(C_0/C)$  was plotted against irradiation time ( $t$ ). The kinetic parameters such as the rate constant, correlation coefficient and substrate half-life were presented in Table 1. Table 1 demonstrated that rate constant increased from  $1.6 \times 10^{-2}$  to  $7.5 \times 10^{-2} \text{ min}^{-1}$  with increasing the initial concentration from 0 to 20 ppm EDTA, respectively.

### 3.1.2 Influence of ZnO dosage

The reduction rate of Cr(VI) was affected by changing of ZnO dosages in presence of 100 ppm EDTA. As depicted in Fig. 2, the photocatalytic reduction of Cr(VI) was rapidly increased with increasing ZnO dosage from 10 to 50 mg. Briefly, it was seen that the reduction of Cr(VI) without ZnO was only 6% after 120 minutes, while 99% reduction was recorded with 40 mg of ZnO dosage under the same reduction condition. No significant result was shown with further accretion of the ZnO dosage. Since the adsorption sites on the ZnO photocatalyst can be increased due to large amount of photocatalyst dosage, hence, the production of photoinduce electrons in conduction band was also increased during this photocatalytic process.

**Table 1** Kinetic parameters for the photocatalytic reduction of Cr(VI)

Concentration of EDTA (ppm)	Rate constant ( $\text{min}^{-1}$ )	$t_{1/2}$ (min)	$R^2$
0	$1.6 \times 10^{-2}$	43	0.97
20	$7.5 \times 10^{-2}$	9	0.99

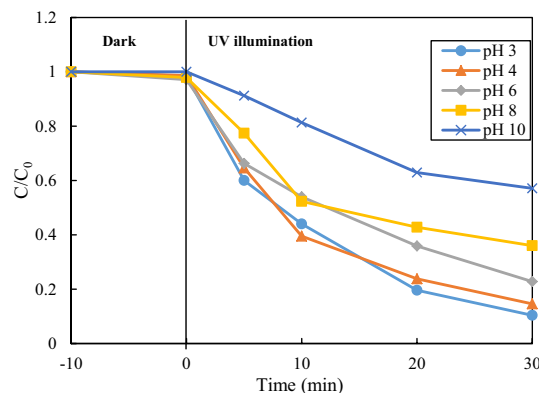


**Fig. 2** Influence of ZnO dosage on the photocatalytic reduction of Cr(VI) in presence of 100 ppm EDTA

Eventually, the reduction efficiency of Cr(VI) would be enhanced greatly [38, 39].

### 3.1.3 Influence of pH

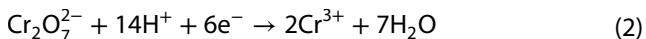
The influence of pH on the photocatalytic reduction is addressed as an important parameter during photocatalytic treatment. The influence of pH was examined with various pH values from 3 to 10 with 40 mg ZnO and 20 ppm EDTA. It is observed in Fig. 3 that the photocatalytic Cr(VI) reduction was raised with declining the initial pH of solution. After 30 min of UV irradiation, the removal efficiency of Cr(VI) was 90%, 85.4%, 77.2%, 64% and 43% at pH=3, 4, 6, 8 and 10, respectively. Therefore, the photocatalytic reduction of Cr(VI) occurred significantly at pH 3 rather than higher pH. This result can be illustrated by the relationships among surface charge of ZnO photocatalyst, pH values and speciation of Cr(VI) in solution [40]. Normally, chromium (VI) species exist as chromate ( $\text{CrO}_4^{2-}$ ), dichromate ( $\text{Cr}_2\text{O}_7^{2-}$ ), hydrogen chromate ( $\text{HCrO}_4^-$ ),



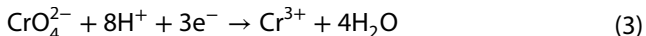
**Fig. 3** Influence of pH on the photocatalytic reduction of Cr(VI) with 20 ppm EDTA

dihydrogen chromate (chromic acid,  $\text{H}_2\text{CrO}_4$ ), hydrogen dichromate ( $\text{HCr}_2\text{O}_7^-$ ), trichromate ( $\text{Cr}_3\text{O}_{10}^{2-}$ ) and tetrachromate ( $\text{Cr}_4\text{O}_{13}^{2-}$ ), where species depends on solution pH. The last three ions ( $\text{HCr}_2\text{O}_7^-$ ,  $\text{Cr}_3\text{O}_{10}^{2-}$  and  $\text{Cr}_4\text{O}_{13}^{2-}$ ) have been observed only in solutions of  $\text{pH} < 0$  or at a chromium (VI) concentration greater than 1 mol/L [41]. At low pH, Cr(VI) exists as  $\text{HCrO}_4^-$  mainly, while  $\text{CrO}_4^{2-}$  occurs at  $\text{pH} > 6.5$  [25]. The  $\text{pH}_{\text{pzc}}$  of ZnO is 9 [42]. Hence, oxyanions of Cr(VI) can be readily absorbed on ZnO surface below the  $\text{pH}_{\text{pzc}}$  since ZnO surface becomes positively charged [43]. Moreover, the thermodynamic driving force between redox potential of Cr(VI)/Cr(III) and conduction band of photocatalyst became narrower with higher pH values as well as less photocatalytic reduction of Cr(VI) could be recorded [44]. The ZnO surface becomes negatively charged under alkaline condition which edges the conversion of  $\text{Cr}_2\text{O}_7^{2-}$  oxyanions to  $\text{CrO}_4^{2-}$  ion as predominant species and lessens the reduction of Cr(VI). Again,  $\text{Cr}(\text{OH})_3$  can be deposited on the photocatalyst surface for decreasing its photocatalytic activity at higher pH [38]. Therefore, pH 3 was finalized as optimal pH for this reduction process. The photocatalytic reduction of  $\text{Cr}^{6+}$  ions to  $\text{Cr}^{3+}$  ions is presented depending on pH through Eqs. (2–4) [45].

At acidic condition:



At neutral condition:

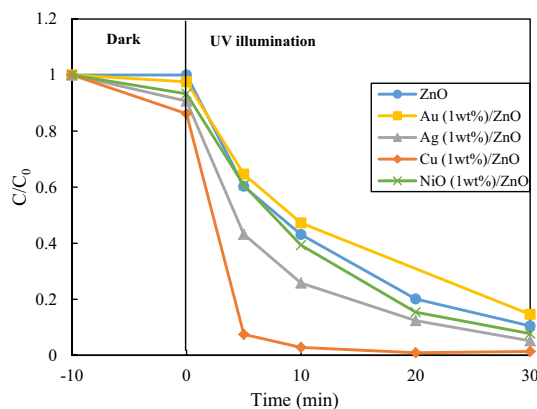


and at alkaline condition:



### 3.1.4 Influence of co-catalyst

A promising method to improve the photocatalytic activity of semiconductor-based photocatalysts is the loading of metal onto semiconductor materials. Thusly, Au, Ag, Cu and NiO were loaded on ZnO photocatalyst during photocatalytic reduction of Cr(VI) to scrutinize the effect of co-catalysts. Here, standard solution of Au, Ag and Cu ions, with given concentrations, were added directly into reaction solution while NiO particles were added by suspension method. The result is shown in Fig. 4. All of the 1 wt% metal modified ZnO resulted higher photocatalytic reduction of Cr(VI) compared with alone ZnO in presence of 20 ppm EDTA. Under UV irradiation, metal ions could trap the excited electrons in conduction band of ZnO photocatalyst and metal ions was reduced to metal and metal oxide which were deposited onto photocatalyst surface. Metal and metal oxide could help to retard the photoinduced electron-hole pair recombination [46]. Therefore, photocatalytic reduction of Cr(VI) was accelerated by 1 wt% metal doped ZnO nanoparticles.

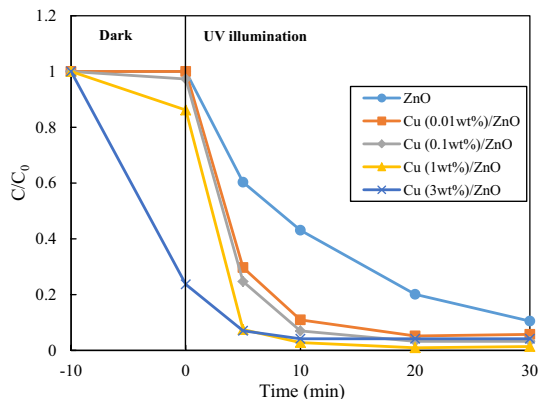


**Fig. 4** Influence of co-catalyst on the reduction of Cr(VI) with ZnO in the presence of EDTA (20 ppm). Additives: Au 10 ppm (1 wt%), Ag 10 ppm (1 wt%), Cu 10 ppm (1 wt%) and NiO 10 ppm (1 wt%)

However, the reduction rate varies insignificantly in the order of  $\text{Cu} > \text{Ag} > \text{NiO} > \text{Au}$ . Although the similar mechanism could be observed for all these metals, photocatalytic activity varied slightly due to metalcatalytic properties. Copper ion could be reduced to both copper metal and  $\text{Cu}_2\text{O}$  by using conduction band electrons of photocatalyst. Cu metal and oxide trapped electron and stopped the electron-hole recombination. Moreover,  $\text{Cu}_2\text{O}$  released the electron and converted into CuO. These electrons could also reduce Cr(VI) to Cr(III). Therefore, photocatalytic reduction of Cr(VI) with Cu motivated ZnO was slightly more than other metals.

### 3.1.5 Influence of Cu amount

Cu was selected as co-catalyst with ZnO during the photocatalytic reduction of Cr(VI) since the photoreduction efficiency with Cu was quite higher compared to other co-catalysts. The effect of concentration of Cu on the photocatalytic Cr(VI) reduction was also studied and explained in Fig. 5. Although the adsorption of Cr(VI) on ZnO with different concentrations of  $\text{Cu}^{2+}$  ion were dissimilar in dark, the reduction efficiency with all concentration of  $\text{Cu}^{2+}$  ion were almost the same after 30 min irradiation. The sequence of reduction efficiency is  $0 \text{ wt}\% < 0.01 \text{ wt}\% < 0.1 \text{ wt}\% < 3 \text{ wt}\% < 1 \text{ wt}\%$ . Hence, 1 wt% Cu accelerated the photocatalytic reduction compared with other concentration. The low doping of Cu may halt the electrons- holes recombination. On contrary, reaction rate can be adversely affected by high doping of Cu. For example, the deposited metallic copper could be oxidized to  $\text{Cu}^{2+}$  ions by consuming valence band holes for 3 wt% Cu doping [35].



**Fig. 5** Influence of Cu doping amount on the reduction of Cr(VI) with ZnO in the presence of EDTA (20 ppm). Additives: Cu 0.1 ppm (0.01 wt %), Cu 1 ppm (0.1 wt%) and Cu 10 ppm (1wt %)

### 3.2 Characterization of ZnO and Cu/ZnO photocatalysts

#### 3.2.1 Textural characterization

The textural properties of ZnO and 1 wt% Cu/ZnO were examined by Brunauer–Emmett–Teller N<sub>2</sub> adsorption–desorption isotherm. The evaluation of the BET surface area and pore distribution of photocatalysts was shown in Fig. 6. As demonstrated in Fig. 6, the adsorption–desorption isotherm of ZnO and Cu/ZnO were confirmed the type IV isotherm, since the hysteresis loop was detected (Fig. 6a inset). The BET surface area, the total pore volume and the average pore diameter were presented in Table 2. All of parameters were increased due to the deposition of Cu with ZnO.

An image obtained by a transmission electron microscopy, is shown in Fig. 7 to study the morphology of ZnO and Cu/ZnO samples. It was clearly observed the nanoparticles were agglomerated and the shape was almost spherical with a hexagonal surface structure for pure ZnO. However, the morphology was considerably changed by

**Table 2** Textural properties of ZnO and 1wt % Cu/ZnO photocatalysts

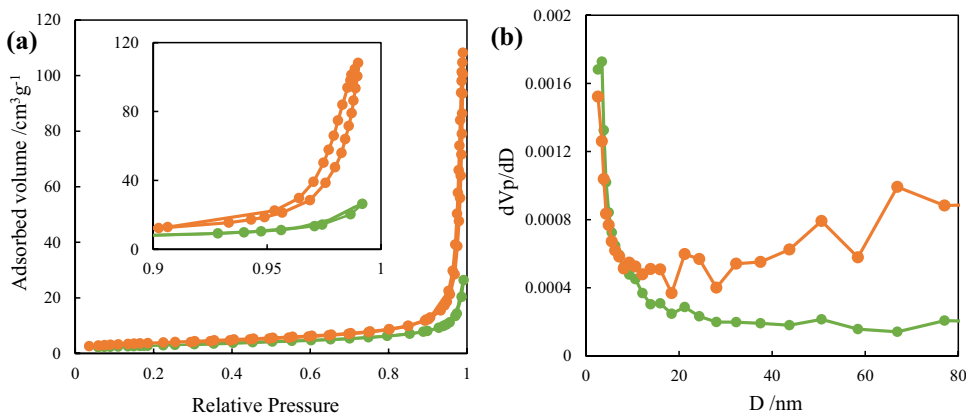
	Surface area (m <sup>2</sup> /g)	Pore volume (cm <sup>3</sup> /g)	Average pore diameter (nm)
ZnO	10.0	0.0374	15.0
Cu(1wt %)/ZnO	13.8	0.167	48.4

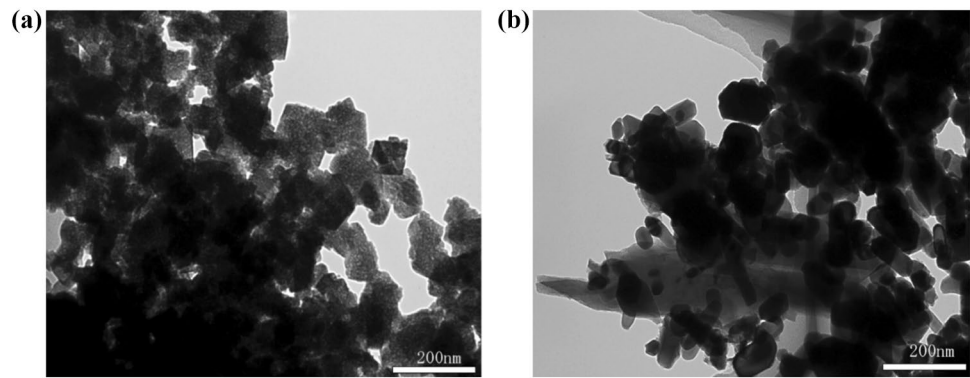
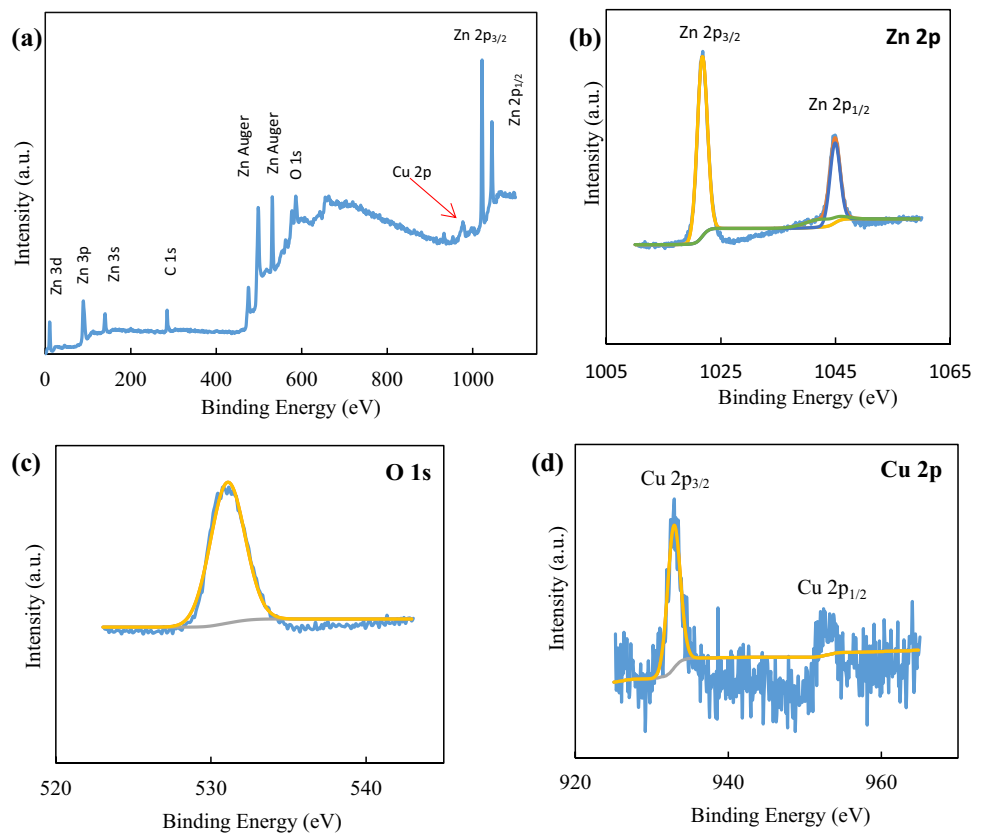
the doping of Cu. The small rod like structure of Cu/ZnO photocatalyst was detected. The Cu doping also suppressed the agglomeration of ZnO. Energy–dispersive X–ray spectroscopy (EDX) was performed for the elemental analysis of ZnO and Cu/ZnO photocatalysts. The electron mapping of Cu/ZnO nanoparticles are presented in Fig. S1 in Supplementary Material. As shown in Fig S1, the presence of Cu with ZnO could be confirmed.

#### 3.2.2 XPS analysis

The chemical bonding state and the surface composition of Cu/ZnO photocatalyst was analyzed by XPS analysis. Figure 8 displays the XPS spectra of Cu/ZnO photocatalyst. The survey XPS spectrum is shown in Fig. 8a, demonstrating that the Zn 2p and Zn Auger peaks in addition to the Cu 2p, C 1s and O 1s peak. From Fig. 8b, the doublet spectral lines of Zn 2p<sub>3/2</sub> and 2p<sub>1/2</sub> due to spin–orbit coupling were recorded at the binding energy of 1021.94 eV and 1045.34 eV, respectively. These values were close to the standard ZnO binding energy values. According to this result, zinc(II) ion (Zn<sup>2+</sup>) was confirmed in Cu/ZnO [3]. The O 1s spectra in Fig. 8c shows one peak located at 531.24 eV, which is attributed to lattice oxygen of ZnO photocatalyst. The high-resolution Cu 2p spectra are shown in Fig. 8d. The peak centered at 933.14 eV is attributed to Cu 2p<sub>3/2</sub>, whereas the peak at 953.74 eV is ascribed to Cu 2p<sub>1/2</sub>, indicating the existence of copper with the +2 oxidation state [47, 48].

**Fig. 6** N<sub>2</sub> adsorption isotherms of ZnO (green) and 1wt % Cu/ZnO (orange) (a); pore size distribution curves of ZnO (green) and 1wt % Cu/ZnO (orange) (b)



**Fig. 7** TEM images of ZnO (a) and Cu/ZnO nanoparticles (b)**Fig. 8** High-resolution XPS spectra for 1 wt% Cu/ZnO nanoparticles. Survey spectra (a); Zn 2p (b); O 1s (c) and Cu 2p (d)

### 3.3 Optical properties

#### 3.3.1 UV-Vis diffuse reflectance spectra

Figure 9 shows a comparison of the UV-Vis diffuse reflection spectrum of ZnO and 1 wt% Cu/ZnO. It was confirmed that Cu/ZnO exhibited a wide visible light absorbance in the range of 400–800 nm as well as the band absorbance of ZnO. The optical band gaps of ZnO and Cu/ZnO nanoparticles were calculated from the UV-Vis diffuse reflectance spectra (Fig. 9). The diffuse reflectance can be converted into absorption coefficient equivalent using Kubelka–Munk function,  $F(R)$ , as represented below [49]

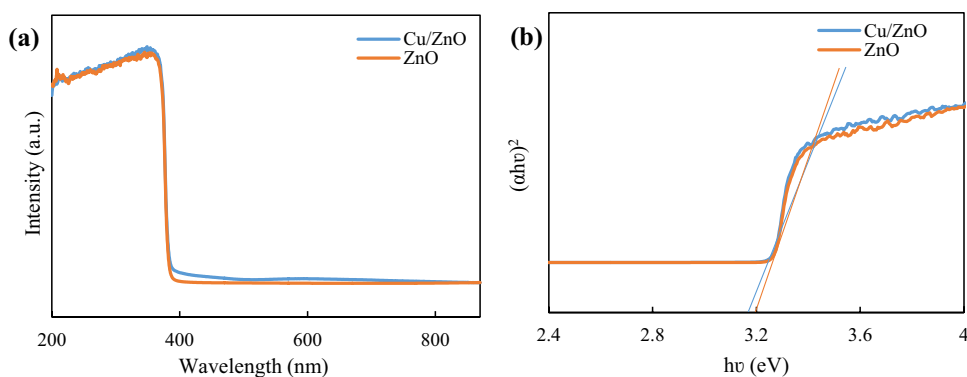
$$F(R) = (1 - R)^2 / 2R \quad (5)$$

where,  $F(R)$  is the Kubelka–Munk function of the sample and  $R$  is the reflectance of the sample. In the parabolic band structure, the  $E_g$  and absorption coefficient are related through the well-known Tauc relation [50]. The Tauc relation for a band gap material is given by the expression

$$\alpha h\nu = A(h\nu - E_g)^n \quad (6)$$

where,  $\alpha$  is linear absorption coefficient,  $h$  is the Planck's constant,  $\nu$  is light frequency,  $A$  is the proportionality constant and  $E_g$  is the band gap energy. The power of the

**Fig. 9** UV-Vis diffuse reflectance spectra of ZnO and 1 wt% Cu/ZnO (a); Tauc plot of ZnO and 1 wt% Cu/ZnO (b)



parenthesis is  $n=2$  or  $\frac{1}{2}$  for indirect and direct band gap materials, respectively. Because direct transition was found for ZnO, the direct band gap ( $E_g$ ) could be obtained by plotting a graph of  $(ahv)^2$  against  $h\nu$  based on the Tauc (Fig. 9b). By extrapolating the linear region of Tauc plot, the band gap of ZnO and 1 wt% Cu/ZnO were found to be 3.21 and 3.18 eV, respectively. The optical band gap of 1 wt% Cu/ZnO was slightly decreased as compared with that of pure ZnO.

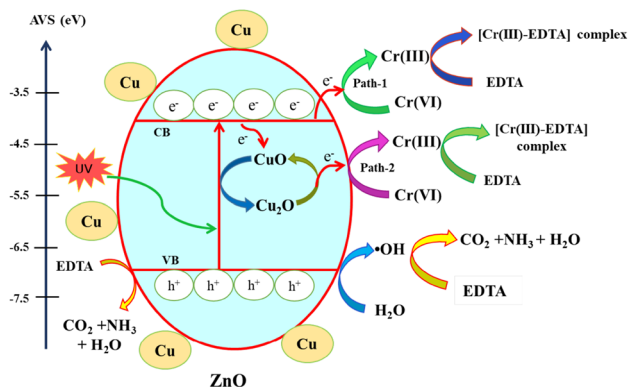
### 3.3.2 Photoluminescence spectra

To elucidate the effect of Cu ion doping on the efficiency of charge-carrier trapping, immigration and transfer, the photoluminescence emission spectra of ZnO and 1 wt% Cu/ZnO with excitation source ranges from 325 to 350 nm are displayed in Fig. S2. Two peaks were detected in the PL spectra of pure ZnO nanoparticles in all of excitation sources. The peak found at about 398 nm in the ultraviolet region may be owing to the recombination of free exciton in the near-band-edge of ZnO. The other peak about 496 nm which exhibited a broad visible emission was generally attributed to the oxygen vacancies in ZnO lattices. Photoluminescence is evidently related with the recombination of the electron-hole pairs. The lower PL intensity confirms the slower photo-generated electron-hole pair recombination showing the enhancement of photocatalytic activity. Here, the PL intensities of Cu/ZnO nanoparticles were significantly lower than that of pure ZnO nanoparticles in all of excitation sources, demonstrating that a better optical properties of Cu/ZnO. The fact could be ascribed to it that the Cu can trap several electrons to suppress the recombination of electron-hole pairs for 1 wt% Cu/ZnO photocatalyst. The comparison of PL intensities explained that Cu/ZnO had a lower recombination rate and better optical properties.

### 3.4 Cr(VI) Reduction mechanism

The mechanism of the photocatalytic reduction of Cr(VI) by 1 wt% Cu/ZnO in presence of EDTA is a very complicated reaction. However a plausible mechanism was presented in Scheme 1. The mechanism was based on three aspects, (1) the small amount of Cu could be acted as electron trapping sites which reduced the electron-hole recombination gently; (2) the one step reduction of Cr(VI) to Cr(III) by the photo-generated electrons ( $e^-$ ); (3) water was oxidized by valence band holes to produce highly active hydroxyl radicals, therefore, sacrificial electron donor (EDTA) could be oxidized by both valence band holes and hydroxyl radicals.

The conduction band and valence band of ZnO are positioned at  $-3.9$  eV and  $-7.1$  eV (vs. AVS), respectively [51]. The estimated band gap of 1 wt% Cu/ZnO was 3.18 eV (Fig. 9). Under UV irradiation, the Cu/ZnO (band-gap of 3.18 eV) absorbed the photons efficiently and was excited to form electrons in the conduction band (CB) and holes in the valence band (VB). Theoretically,  $Cu^{2+}$  ions are reduced to Cu metal and  $Cu_2O$ , which are deposited on ZnO surface in solution under UV irradiation [52]. The standard reduction potential of  $Cu^{2+}/Cu$  and  $Cu^{2+}/Cu^+$  are  $+0.35$  V and  $+0.15$  V versus NHE, respectively [53,

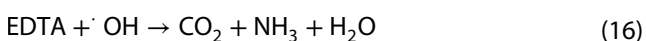
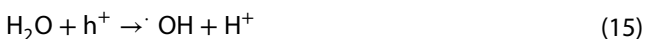
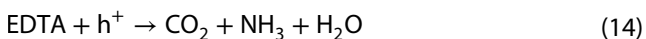
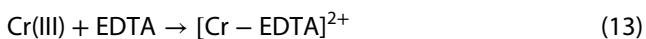
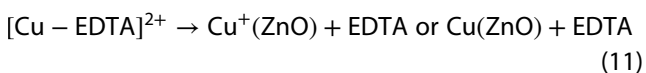
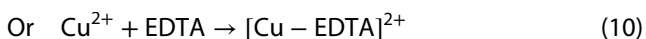


**Scheme 1** Proposed mechanism for photocatalytic Cr(VI) reduction with 1 wt% Cu/ZnO in presence of EDTA under UV illumination

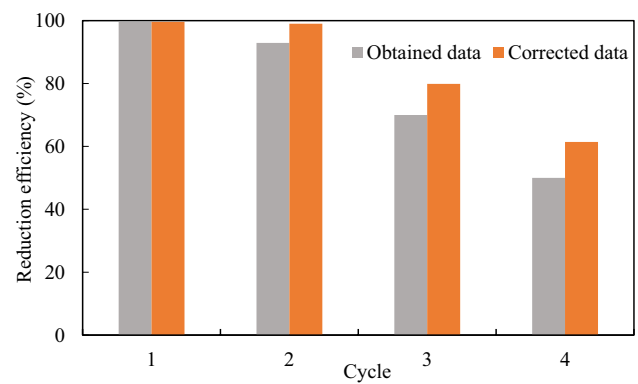


54]. Hence, reduction of  $\text{Cu}^{2+}$  ion to metallic Cu and  $\text{Cu}^+$  ion are thermodynamically possible by photogenerated electron of less negative ZnO conduction band potential. Here,  $\text{Cu}^+$  ions were quickly converted into  $\text{Cu}_2\text{O}$ , since  $\text{Cu}^+$  ions were quite unstable. The released electron from  $\text{Cu}_2\text{O}$  could reduce Cr(VI) to Cr(III). In addition, the Cr(VI)/Cr(III) redox couple has a more positive reduction potential,  $E^0(\text{HCrO}_4^-/\text{Cr}^{3+}) = +1.35 \text{ V}$  versus NHE [55] with respect to ZnO conduction band edge. Therefore, the originated conduction band electrons reduce Cr(VI) to Cr(III) thermodynamically. EDTA has been applied to enhance the photocatalytic Cr(VI) reduction with Cu/ZnO. The complex of EDTA with reduced Cr(III),  $[\text{Cr}-(\text{EDTA})]^{3+}$  as final product of this reduction process.  $[\text{Cr}-(\text{EDTA})]^{3+}$  complex is soluble and cannot occur on ZnO surface. Therefore, reduction process was benefited.

The point of zero charge (pzc) of the ZnO particle is approximately 9. When the pH is lower than this value, the ZnO surface becomes positively charged. From the estimation of Cr(VI) species as the function of pH, the main chemical species for hexavalent chromium at pH 3 is  $\text{HCrO}_4^-$ . Owing to the electrostatic attraction between  $\text{HCrO}_4^-$  species and ZnO photocatalyst with a relatively large surface area, as well as the facilitation of the proton under acidic conditions, adsorbed Cr(VI) on the surface of ZnO can be immediately reacted with electrons and be reduced to Cr(III). The reaction related in this photoreduction process are represented in Eqs. (7–16).



The deactivation of photocatalyst surface, due to adsorption of Cr(III) species, could be prevented by the formation of  $[\text{Cr}-(\text{EDTA})]^{3+}$  complex in reaction solution [36, 56]. Simultaneously, EDTA could consume valence band holes ( $h^+$ ) directly, and EDTA was oxidized by hydroxyl radical. EDTA was converted to carbon dioxide, ammonia and



**Fig. 10** Reusability of 1 wt% Cu/ZnO for the reduction of Cr(VI) water after several steps. Thus, EDTA played dual role during this reduction process. As a result, the recombination of electrons and holes was suppressed favoring the fast photocatalytic Cr(VI) reduction.

### 3.5 Reusability

The reusability of 1 wt% Cu/ZnO was investigated by conducting the circulating runs in the photocatalytic Cr(VI) reduction under UV illumination. The reaction solution was separated after completing one cycle, therefore, a certain amount of photocatalyst could be lost after every cycle. Hence, the obtained reduction efficiency should be corrected, taking account of the decreased amount of photocatalyst in the reaction solution. As shown in Fig. 10, the photocatalytic activity of Cu/ZnO was decreased slowly with every cycles and after fourth cycles about 60% of reduction efficiency was recorded.

## 4 Conclusion

In summary, this research work aids a procedure to enrich the photocatalytic activity of ZnO nanoparticles with 1 wt% Cu doping in presence of ethylene diaminetetraacetic acid (EDTA) to reduce Cr(VI) to Cr(III). The effect of factors revealed that the small particle size of ZnO, low pH value ( $\text{pH} > 3$ ), the proper amount EDTA and Cu doping and ZnO, photocatalyst dosage could promote the reduction of Cr(VI). The highest 98.5% of reduction of Cr(VI) was recorded with 20 ppm EDTA after only 120 min. Increasing the photocatalyst dosages resulted the significant improvement in the reduction of Cr(VI). Similarly, photocatalytic reduction of Cr(VI) led to decreasing from 90 to 43% as initial pH increased 3 to 10 after 30 min UV irradiation. Briefly, Cu doping retarded the photoinduced electron-hole recombination while EDTA immensely influenced this photocatalytic Cr(VI) reduction process by removing of Cr(III) from ZnO photocatalyst surface

and capturing holes of valence band. These two factors could simultaneously provide a sovereign way for the efficient removal of Cr(VI) under UV irradiation. Therefore, the present treatment method is simple, low cost and easy-handling.

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## Compliance with ethical standards

**Conflict of interest** The authors declare no conflict of interest.

## References

1. Korak JA, Huggins R, Arias-Paic P (2017) Regeneration of pilot-scale ion exchange columns for hexavalent chromium removal. *Water Res* 118:141–151. <https://doi.org/10.1016/j.watres.2017.03.018>
2. Du XD, Yi XH, Wang P, Zheng W, Deng J, Wang CC (2019) Robust photocatalytic reduction of Cr(VI) on UiO-66-NH<sub>2</sub>(Zr/Hf) metal-organic framework membrane under sunlight irradiation. *Chem Eng J* 356:393–399. <https://doi.org/10.1016/j.cej.2018.09.084>
3. Bao C, Chen M, Jin X, Hu D, Huang Q (2019) Efficient and stable photocatalytic reduction of aqueous hexavalent chromium ions by polyaniline surface-hybridized ZnO nanosheets. *J Mol Liq* 279:133–145. <https://doi.org/10.1016/j.molliq.2019.01.122>
4. Bakshi A, Panigrahi AK (2018) A comprehensive review on chromium induced alterations in fresh water fishes. *Toxicol Rep* 5:440–447. <https://doi.org/10.1016/j.toxrep.2018.03.007>
5. Shahid M, Shamshad S, Rafiq M, Khalid S, Bibi I, Niazi NK, Dumat C, Rashid MI (2017) Chromium speciation, bioavailability, uptake, toxicity and detoxification in soil-plant system: a review. *Chemosphere* 178:513–533. <https://doi.org/10.1016/j.chemosphere.2017.03.074>
6. Li J, Liu Z, Zhu Z (2015) Enhanced photocatalytic activity of ZnFe<sub>2</sub>O<sub>4</sub>-ZnO-Ag<sub>3</sub>O<sub>4</sub> hollow nanospheres through the cascaded electron transfer with magnetical separation. *J Alloys Compd* 636:229–233. <https://doi.org/10.1016/j.jallcom.2015.02.176>
7. Samadi MT, Saghi MH, Rahmani A, Rahimi S, Syboney MS (2010) Hamadan landfill leachate treatment by coagulation-flocculation process. *Iran J Environ Health Sci Eng* 7(3):253–258
8. Siboni MS, Khataee A, Vafaei F, Joo SW (2014) Comparative removal of two textile dyes from aqueous solution by adsorption onto marine-source waste shell: kinetic and isotherm studies. *Korean J Chem Eng* 31(8):1451–1459. <https://doi.org/10.1007/s11814-014-0085-4>
9. Siboni MS, Jafari S-J, Farrokhi M, Yang JK (2013) Removal of phenol from aqueous solutions by activated red mud: equilibrium and kinetics studies. *Environ Eng Res* 18(4):247–252. <https://doi.org/10.4491/eer.2013.18.4.247>
10. Natale FD, Erto A, Lancia A, Musmarra D (2015) Equilibrium and dynamic study on hexavalent chromium adsorption onto activated carbon. *J Hazard Mater* 281:47–55. <https://doi.org/10.1016/j.jhazmat.2014.07.072>
11. Qi H, Wang S, Liu H, Gao Y, Wang T, Huang Y (2016) Synthesis of an organic-inorganic polypyrrole/titanium(IV) biphosphate hybrid for Cr(VI) removal. *J Mol Liq* 215:402–409. <https://doi.org/10.1016/j.molliq.2015.12.060>
12. Mohagheghian A, Karimi SA, Yang JK, Siboni MS (2015) Photocatalytic degradation of a textile dye by illuminated tungsten oxide nanopowder. *J Adv Oxid Technol* 18(1):61–68. <https://doi.org/10.1515/jaots-2015-0108>
13. Saien J, Azizi A, Soleymani AR (2014) Parameter evaluation, kinetics, and energy consumption for Cr(VI) photocatalytic reduction under mild conditions. *J Iran Chem Soc* 11:1439–1448. <https://doi.org/10.1007/s13738-014-0419-5>
14. Yan X, Ning G, Zhao P (2019) Enhanced visible light photocatalytic reduction of Cr(VI) over a novel square nanotube poly(triazine amide)/TiO<sub>2</sub> heterojunction. *Catalysts* 9:55–66. <https://doi.org/10.3390/catal9010055>
15. Ravishankar TN, Vaz MO, Ramakrishnappa T, Teixeira SR, Dupont J (2017) Ionic liquid assisted hydrothermal syntheses of Au doped TiO<sub>2</sub> NPs for efficient visible-light photocatalytic hydrogen production from water, electrochemical detection and photochemical detoxification of hexavalent chromium (Cr<sup>6+</sup>). *RSC Adv* 7:43233–43244. <https://doi.org/10.1039/C7RA04944G>
16. Ibhaddon AO, Fitzpatrick P (2013) Heterogeneous photocatalysis: recent advances and applications. *Catalysts* 3(1):189–218. <https://doi.org/10.3390/catal3010189>
17. Siboni MS, Samarghandi MR, Yang JK, Lee SM (2011) Photocatalytic removal of cyanide with illuminated TiO<sub>2</sub>. *Wat Sci Tech* 64(7):1383–1387. <https://doi.org/10.2166/wst.2011.738>
18. Mohagheghian A, Karimi SA, Yang JK, Siboni MS (2016) Photocatalytic degradation of diazinon by illuminated WO<sub>3</sub> nanopowder. *Desalin Water Treat* 57(18):8262–8269. <https://doi.org/10.1080/19443994.2015.1017009>
19. Mohagheghian A, Pourmohseni M, Vahidi-Kolur R, Yang JK, Siboni MS (2017) Application of kaolin-Fe<sub>3</sub>O<sub>4</sub> nano-composite for the removal of azo dye from aqueous solutions. *Desalin Water Treat* 58:308–319. <https://doi.org/10.5004/dwt.2017.0198>
20. Humayun M, Raziq F, Khan A, Luo W (2018) Modification strategies of TiO<sub>2</sub> for potential applications in photocatalysis: a critical review. *Green Chem Lett Rev* 11(2):86–102. <https://doi.org/10.1080/17518253.2018.1440324>
21. Guo Q, Zhou C, Ma Z, Ren Z, Fan H, Yang X (2016) Elementary photocatalytic chemistry on TiO<sub>2</sub> surfaces. *Chem Soc Rev* 45:3701–3730. <https://doi.org/10.1039/C5CS00448A>
22. Siboni MS, Khataee A, Vahid B, Joo SW (2015) Synthesis, characterization and immobilization of ZnO nanosheets on scallop shell for photocatalytic degradation of an insecticide. *Sci Adv* 7(4):806–814. <https://doi.org/10.1166/sam.2015.2163>
23. Siboni MS, Khataee A, Vahid B, Joo SW, Fallah S (2014) Preparation of a green photocatalyst by immobilization of synthesized ZnO nanosheets on scallop shell for degradation of an azo dye. *Curr Nanosci* 10(5):684–694. <https://doi.org/10.2174/1573413710666140521175816>
24. Chu DW, Masuda Y, Ohji T, Kato K (2010) Formation and photocatalytic application of ZnO nanotubes using aqueous solution. *Langmuir* 26:2811–2815. <https://doi.org/10.1021/la902866a>
25. Islam JB, Furukawa M, Tateshi I, Katsumata H, Kaneco S (2019) Photocatalytic reduction of hexavalent chromium with nano-sized TiO<sub>2</sub> in presence of formic acid. *ChemEngineering* 3:33–42. <https://doi.org/10.3390/chemengineering3020033>
26. Rameshbabu R, Kumar N, Karthigeyan A, Neppolian B (2016) Visible light photocatalytic activities of ZnFe<sub>2</sub>O<sub>4</sub>/ZnO nanoparticles for the degradation of organic pollutants. *Mater Chem Phys* 18:106–115. <https://doi.org/10.1016/j.matchemphys.2016.06.040>
27. Gholami M, Siboni MS, Farzadkia M, Yang JK (2016) Synthesis, characterization, and application of ZnO/TiO<sub>2</sub> nanocomposite for photocatalysis of a herbicide (Bentazon). *Desalin Water Treat* 57(29):13632–13644. <https://doi.org/10.1080/19443994.2015.1060541>

28. Siboni MS, Jonidi-Jafari A, Farzadkia M, Esrafil A, Gholami M (2017) Enhancement of photocatalytic activity of Cu-doped ZnO nanorods for the degradation of an insecticide: kinetics and reaction pathways. *J Environ Manage* 186:1–11. <https://doi.org/10.1016/j.jenvman.2016.10.049>
29. Gholami M, Siboni MS, Yang JK (2016) Application of Ni-doped ZnO rods for the degradation of an azo dye from aqueous solutions. *Korean J Chem Eng* 33(3):812–882. <https://doi.org/10.1007/s11814-015-0218-4>
30. Samarghandi MR, Yang JK, Lee SM, Giahi O, Siboni MS (2014) Effect of different type of organic compounds on the photocatalytic reduction of Cr(VI) in presence of ZnO nanoparticles. *Desalin Water Treat* 52(7–9):1531–1538. <https://doi.org/10.1080/19443994.2013.797624>
31. Babay PA, Emilio CA, Ferreyra RE, Gautier EA, Gettar RT, Litter MI (2001) Kinetics and mechanisms of EDTA photocatalytic degradation with TiO<sub>2</sub>. *Wat Sci Tech* 44(5):179–185. <https://doi.org/10.2166/wst.2001.0281>
32. Westin KJ, Rasmuso AC (2005) Nucleation of calcium carbonate in presence of citric acid, DTPA, EDTA and pyromellitic acid. *J Colloid Interf Sci* 282:370–379. <https://doi.org/10.1016/j.jcis.2004.09.071>
33. Siemon U, Bahnemann D, Testa JJ, Rodríguez D, Litter MI, Bruno N (2002) Heterogeneous photocatalytic reactions comparing TiO<sub>2</sub> and Pt/TiO<sub>2</sub>. *J Photochem Photobiol, A* 148:247–255. [https://doi.org/10.1016/S1010-6030\(02\)00050-3](https://doi.org/10.1016/S1010-6030(02)00050-3)
34. Bhanushali S, Ghosh P, Ganesh A, Cheng WL (2015) 1D copper nanostructures: progress, challenges and opportunities. *Small* 11:1232–1252. <https://doi.org/10.1002/sml.201402295>
35. Janczarek M, Kowalska E (2017) On the origin of enhanced photocatalytic activity of copper-modified titania in the oxidative reaction systems. *Catalysts* 7:317–343. <https://doi.org/10.3390/catal7110317>
36. Meichtry JM, Colbeau-Justin C, Custo G, Litter MI (2014) Preservation of the photocatalytic activity of TiO<sub>2</sub> by EDTA in the reductive transformation of Cr(VI). *Studies by time resolved microwave conductivity. Catal Today* 224:236–243. <https://doi.org/10.1016/j.cattod.2013.10.021>
37. Sofi FA, Majid K, Mehraj O (2018) The visible light driven copper based metal-organic-framework heterojunction: HKUST-1@Ag-Ag<sub>3</sub>PO<sub>4</sub> for plasmon enhanced visible light photocatalysis. *J Alloys Compd* 737:798–808. <https://doi.org/10.1016/j.jallcom.2017.12.141>
38. Shao D, Wang X, Fan Q (2009) Photocatalytic reduction of Cr(VI) to Cr(III) in solution containing ZnO or ZSM-5 zeolite using oxalate as model organic compound in environment. *Microporous Mesoporous Mater* 117:243–248. <https://doi.org/10.1016/j.micromeso.2008.06.026>
39. Ku Y, Jung IL (2001) Photocatalytic reduction of Cr(VI) in aqueous solutions by UV irradiation with the presence of titanium dioxide. *Water Res* 35:135–142. [https://doi.org/10.1016/S0043-1354\(00\)00098-1](https://doi.org/10.1016/S0043-1354(00)00098-1)
40. Xu Q, Li R, Wang C, Yuan D (2017) Visible-light photocatalytic reduction of Cr(VI) using nano-sized delafossite (CuFeO<sub>2</sub>) synthesized by hydrothermal method. *J Alloys Compd* 723:441–447. <https://doi.org/10.1016/j.jallcom.2017.06.243>
41. Szabó M, Kalmár J, Ditrói T, Bellér G, Lente G, Simic N, Fábrián I (2018) Equilibria and kinetics of chromium(VI) speciation in aqueous solution—a comprehensive study from pH 2 to 11. *Inorg Chim Acta* 472:295–301. <https://doi.org/10.1016/j.ica.2017.05.038>
42. Naimi-Joubani M, Shirzad MS, Yang JK, Gholami M, Farzadkia M (2015) Photocatalytic reduction of hexavalent chromium with illuminated ZnO/TiO<sub>2</sub> composite. *J Ind Eng Chem* 22:317–323. <https://doi.org/10.1016/j.jiec.2014.07.025>
43. Liu SX (2005) Removal of copper (VI) from aqueous solution by Ag/TiO<sub>2</sub> photocatalysis. *Bull Environ Contam Toxicol* 74:706–714. <https://doi.org/10.1007/s00128-005-0640-0>
44. Wang N, Zhu L, Deng K, She Y, Yu Y, Tang H (2010) Visible light photocatalytic reduction of Cr(VI) on TiO<sub>2</sub> on situ modified with the small molecular weight organic acids. *Appl Catal B-Environ* 95(3–4):400–407. <https://doi.org/10.1016/j.apcatb.2010.01.019>
45. Shirzad MS, Samadi MT, Yang JK, Lee SM (2012) Photocatalytic removal of Cr(VI) and Ni(II) by UV/TiO<sub>2</sub>: kinetic study. *Desalin Wat Treat* 40:77–83. <https://doi.org/10.1080/19443994.2012.671144>
46. Wu SX, Ma Z, Qin YN, He F, Jia LS, Zhang YJ (2003) XPS study of copper doping TiO<sub>2</sub> photocatalyst. *Acta Phys Chim Sin* 19:967–969. <https://doi.org/10.3866/PKU.WHXB20031017>
47. Pashchanka M, Hoffmann RC, Gurlo A, Swarbrick JC, Khanderi J, Engstler J, Issanin A, Schneider JJ (2011) A molecular approach to Cu doped ZnO nanorods with tunable dopant content. *Dalton Trans* 40:4307–4314. <https://doi.org/10.1039/c0dt01567a>
48. Pung S, Ong C, Isha KM, Othman M (2014) Synthesis and characterization of Cu-doped ZnO nanorods. *Sains Malays* 43:273–281
49. Das TK, Illaiyara P, Mocherla PSV, Bhalerao GM (2016) Influence of surface disorder, oxygen defects and bandgap in TiO<sub>2</sub> nanostructures on the photovoltaic properties of dye sensitized solar cells. *Sol Energy Mater Sol Cells* 144:194–209. <https://doi.org/10.1016/j.solmat.2015.08.036>
50. Valenzuela MA, Bosch P, Jiménez-Becerrill J, Quiroz O, Páez AI (2002) Preparation, characterization and photocatalytic activity of ZnO, Fe<sub>2</sub>O<sub>3</sub> and ZnFe<sub>2</sub>O<sub>4</sub>. *J Photochem Photobiol, A* 148:177–182. [https://doi.org/10.1016/S1010-6030\(02\)00040-0](https://doi.org/10.1016/S1010-6030(02)00040-0)
51. Tahir M, Tahir B, Amina NAS, Alias H (2016) Selective photocatalytic reduction of CO<sub>2</sub> by H<sub>2</sub>O/H<sub>2</sub> to CH<sub>4</sub> and CH<sub>3</sub>OH over Cu-promoted In<sub>2</sub>O<sub>3</sub>/TiO<sub>2</sub> nanocatalyst. *Appl Surf Sci* 389:46–55. <https://doi.org/10.1016/j.apsusc.2016.06.155>
52. Jacobs JWM, Kampers FWH, Rikken JMG, Bulle-Lieuwma CWT, Koningsberger DC (1989) Copper photodeposition on titanium dioxide studied with HREM and EXAFS. *J Electrochem Soc* 136:2914–2923. <https://doi.org/10.1149/1.2096373>
53. Sykora J (1997) Photochemistry of copper complexes and their environmental aspects. *Coord Chem Rev* 159:95–105. [https://doi.org/10.1016/S0010-8545\(96\)01299-4](https://doi.org/10.1016/S0010-8545(96)01299-4)
54. Wang L, Lia X, Teng W, Zhao Q, Shi Y, Yue R, Chen Y (2013) Efficient photocatalytic reduction of aqueous Cr(VI) over flower-like SnIn<sub>4</sub>S<sub>8</sub> microspheres under visible light illumination. *J Hazard Mater* 244–245:681–688. <https://doi.org/10.1016/j.jhazmat.2012.10.062>
55. Zhu J, Shao C, Li X, Han C, Yang S, Ma J, Li X, Liu Y (2018) Immobilization of ZnO/polyaniline heterojunction on electrospun polyacrylonitrile nanofibers and enhanced photocatalytic activity. *Mater Chem Phys* 214:507–515. <https://doi.org/10.1016/j.matchemphys.2018.04.053>
56. Choi Y, Koo MS, Bokare AD, Kim D, Bahnemann DW, Choi W (2017) Sequential process combination of photocatalytic oxidation and dark reduction for the removal of organic pollutants and Cr(VI) using Ag/TiO<sub>2</sub>. *Environ Sci Technol* 51:3973–3981. <https://doi.org/10.1021/acs.est.6b06303>

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